



## **Microbial growth yield estimates from thermodynamics and its importance for degradation of pesticides and formation of biogenic non-extractable residues**

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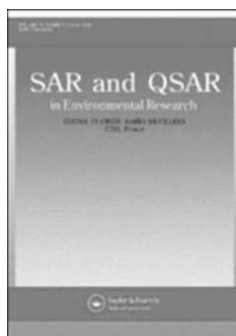
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Keywords:	xenobiotics, biodegradation, microbial biomass, turnover modelling, bound residues, organic chemicals of environmental concern

**1 Microbial growth yield estimates from thermodynamics and its**  
**2 importance for degradation of pesticides and formation of biogenic**  
**3 non-extractable residues**

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**Microbial growth yield estimates from thermodynamics and its importance for degradation of pesticides and formation of biogenic non-extractable residues**

In biodegradation studies with isotope-labelled pesticides fractions of non-extractable residues (NER) remains, but their nature and composition is rarely known, leading to uncertainty about their risk. Microbial growth leads to incorporation of carbon into the microbial mass, resulting in biogenic NER. Formation of microbial mass can be estimated from the microbial growth yield but experimental data is rare. Instead, we suggest using prediction methods for the theoretical yield based on thermodynamics. Recently, we suggested the Microbial Turnover to Biomass (MTB) method that needs minimum input data. We have estimated the growth yield on 40 organic chemicals (31 pesticides) using the MTB and two existing methods. The results were compared to experimental values, and the sensitivity of the methods was assessed. The MTB method performed best for pesticides. Having the theoretical yield and using the released CO<sub>2</sub> as a measure for microbial activity, we predicted a range for the formation of biogenic NER. For the majority of the pesticides, a considerable fraction of the NER was estimated to be biogenic. This novel approach provides a theoretical foundation applicable to evaluate and predict biogenic NER formation during pesticide degradation experiments and may also be employed to interpret NER data from regulatory studies.

Keywords: xenobiotics, biodegradation, microbial biomass, turnover modelling, bound residues, organic chemicals of environmental concern

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2  
3 39 **1. Introduction**

4 40 The evaluation of biodegradation of organic chemicals of environmental concern is a  
5  
6  
7 41 big challenge for risk assessment and is subject to legislation and regulation. In the  
8  
9 42 European Union (EU) chemicals that are traded have to be approved by the REACH  
10  
11 43 legislation if produced and sold in amounts greater than one ton per year [1-3]. The  
12  
13 44 assessment of biodegradability under environmental conditions is standardised by  
14  
15  
16 45 OECD testing guidelines, such as OECD Tests Nos. 306-309 used for the assessment of  
17  
18 46 biodegradation in sea water, soil, fresh water, and fresh-water-sediment systems [4-7].

19  
20 47 Transformation and biodegradation is mostly tested with  $^{14}\text{C}$  or  $^{13}\text{C}$  labelled  
21  
22 48 parent compounds. Isotopes are particularly needed for assessment of non-extractable  
23  
24 49 residues (NER; also called “bound residues”) and tracing of unknown metabolites [8].  
25  
26 50 Assessment of biodegradation is well established but may still have some pitfalls for  
27  
28 51 various compounds [9-12]. Although there are several approaches for the reliable  
29  
30 52 prospective assessment of chemical properties and behaviour from quantitative-  
31  
32 53 structure-activity-relationship (QSAR) modelling available [13], much less is available  
33  
34 54 for the assessment of biodegradation [14]. The assessment of residue formation is still  
35  
36 55 in its infancy and and is not yet predictable.  
37  
38  
39

40 56 Recently, a novel approach was suggested for modelling the formation of  
41  
42 57 biogenic residues [15] which can also elucidate the black box of NER. NER may be  
43  
44 58 formed by sequestration or entrapment of parent compounds or metabolites in soils and  
45  
46 59 sediments (*type I NER*), and also by covalent bonding to soil organic matter (*type II*  
47  
48 60 NER). Another type of residues is formed after incorporation of carbon into microbial  
49  
50 61 biomass after microbial productive degradation of the parent compound (*type III =*  
51  
52 62 *biogenic NER*) [8]. Apparently, NER are mostly comprised of all types of residues and  
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thus the assessment of the biogenic NER formation will also provide information about the amounts of the other types of NER formed.

Compounds that are poor growth substrates and do not provide sufficient energy, carbon or nutrients to the degrader microbes give no incentive for degradation and can be expected to be more stable under environmental conditions. Thus, the usability of the molecule, its energy content and suitability for anabolic processes has a profound impact on the evolutionary pressure to develop degradation pathways.

Thermodynamic analysis can be applied to determine the feasibility and direction of chemical reactions under a given set of conditions (see, e.g., [16]). In addition, thermodynamics can also be used to describe the potential growth of bacteria [17]. Essentially, bacterial growth is simplified and split into anabolic processes (energy demanding) and catabolic processes (energy producing) [18]. The catabolic processes describe the energy released from the oxidation of a chemical or a substrate. In aerobic metabolism the oxidation product is usually  $\text{CO}_2$  and  $\text{H}_2\text{O}$ . The electrons derived from the oxidation are partly transferred to the synthesis of microbial biomass (anabolism) or transferred to the terminal electron acceptors via the respiration chains (catabolism) in the membranes of the organisms resulting in the release of energy, predominately from the formation of  $\text{CO}_2$  and  $\text{H}_2\text{O}$  [19].

The anabolic processes describe the substrate and energy use for the synthesis of new cell biomass. Element, electron and energy balances are used to describe the processes related to the oxidation and reduction half-reactions of the catabolism and to the energy and electron gain used for cell synthesis (see, e.g., [17, 20-22]).

The bacterial growth yield is defined as the mass of bacteria formed per mass of substrate consumed (g cells per g substrate, often also g C per g C) [19].

Thermodynamic growth-yield estimation methods were developed in order to model

metabolic processes and to estimate the amounts of biomass that can be derived from metabolism (e.g. [15, 17, 20-25]). These estimates have previously been used for biotechnological purposes and for the estimation of activated sludge formation in waste water treatment processes, e.g. [26]. The different growth yield estimation methods are based on a similar set of considerations [26, 27].

The Thermodynamic Electron Equivalent Model (TEEM2) developed by Perry L. McCarty [17, 23] and Expanded Thermodynamic True Yield Prediction Model (ETTYM) of Xiao and VanBriesen [22, 24] or their variations have been applied [28, 29] for the estimation of bacterial growth yield on xenobiotics. Both models have evolved towards an increased need of knowledge regarding the transformation pathways, metabolic processes and the electron and energy losses associated hereto in order to model specific growth of various organisms. Recently, the Microbial Turnover to Biomass (MTB) method was developed by Trapp *et al.* [15] in order to provide a less complex method for the estimation of biomass formation during metabolism of any organic compound. In MTB we proposed a simple method to predict just the minimum bacterial growth yield potential without the need for information on the pathways, as this is rarely known for the majority of chemicals of environmental concern. Furthermore, microbial growth and decline are coupled to the formation of soil organic matter (SOM) [30-33]. Therefore, yield estimation provides a tool for the assessment and prediction of biogenic NER formation in the degradation assessment of chemicals for regulatory purposes.

The objectives of the present study are i) to thoroughly compare the recently introduced MTB method with other yield estimation methods; ii) to extend the MTB approach for electron acceptors other than oxygen (e.g. nitrate and sulphate); iii) to predict the biomass yields during pesticide degradation and eventually the potential for

biogenic NER formation; and iv) to determine the sensitive parameters in the yield prediction methods. The estimates were compared to experimentally determined growth yields. Finally, we contrast measured NER formation of pesticides and xenobiotics with the predicted formation of SOM due to bacterial growth and decline.

## 2. Materials and Methods

The bacterial growth yield prediction methods chosen for this study have a common basic approach: a stoichiometrically balanced redox reaction and the associated change of Gibbs free energy. This means that one can set up half-reactions describing the reduction of the targeted compound (be it xenobiotic or not), calculate the associated Gibbs free energy [34], and combine it with half-reactions of an appropriate electron acceptor (e.g.  $O_2$ ,  $NO_3^-$ ,  $SO_4^{2-}$ ,  $Fe^{3+}$ ,  $Mn^{4+}$ , or even  $CO_2$  etc.) allowing for assessment of the bacterial growth yield under a multitude of redox conditions. Here we only look at  $O_2$ ,  $NO_3^-$ , and  $SO_4^{2-}$ . This approach has been shown for ETTYM and TEEM2 and here we will show that the procedure of half-reactions is applicable also for the MTB method.

A detailed summary of the methods can be found in the Supplementary Information (SI) and in the original references. The element, energy and electron balances differ between the methods, thus a brief outline of the methods will be given in Table 1, in which the final equations used to calculate the growth yield are shown.

### 2.1 Microbial Turnover to Biomass (MTB)

The Microbial Turnover to Biomass (MTB) method is presented in detail in Trapp *et al.* [15]. The method is based on the work of Diekert [18]. The maximum bacterial yield is determined from the nutritional value of the substrate (N) combined with the determination of bio-available electrons from the reaction. The nutritional



value is the inverse of the yield and describes how much substrate is needed for the growth of bacteria [g substrate (g biomass)<sup>-1</sup>]. This is subdivided into a biomass yielding (anabolic) and energy yielding (catabolic) part. The catabolic yield is determined from calculation of the Gibbs free energy released from the oxidation of the compound, the storage of this energy in ATP, and the bacterial growth yield on ATP. Microbes cannot use all electrons to generate energy and thus the concept of bio-available electrons was introduced. [15]. Thus, energy and electron balances are implicitly considered. The anabolic yield is calculated from the carbon content in the compound (the carbon source) and in the bacterial cell [18], i.e. how many grams of cell can be produced from the carbon in the compound (only carbon availability is assumed to limit growth). Further details and examples can be found in SI 1.1 and Trapp *et al.* [15].

## 2.2 Thermodynamic Electron Equivalent Model 2 (TEEM2)

In 1965 P. L. McCarty presented a thermodynamic model to estimate the maximal bacterial yield from a single substrate [17]. The method determines the yield on a given substrate from the Gibbs free energy released in the redox process. Since its inception it has been modified and expanded [34]. It was recently modified to better capture the observed lower yields associated with C1 compounds (i.e. methanol) and reactions involving oxygenases [22, 23]. It is based on electron and energy balances. The electron balance considers that the electrons provided by the substrate are used either in the synthesis of cell material (anabolism) or in energy generation (catabolism), and the energy balance states that the energy *captured* with a specific efficiency ( $\epsilon$ ) by the organism is used for bacterial growth. The energy capture efficiency,  $\epsilon$ , is a key

parameter and is estimated from experimental data. Further details and examples are found in SI 1.2 and McCarty [23] and Rittmann and McCarty [34].

### 2.3 Expanded Thermodynamic True Yield Prediction Model (ETTYM)

The Expanded Thermodynamic True Yield Model is based on the work by McCarty and was presented in [22] and expanded in [24]. To increase the accuracy for the yield prediction on C1 compounds and substrates with low degrees of reduction, the authors proposed to include a carbon and a nitrogen balance and as a result reformulated the electron and energy balance originally proposed by McCarty. The carbon balance describes the carbon as either invested in cell synthesis or into other carbonaceous products. The nitrogen balance can be ignored if nitrogen is not limiting [24], hence, the yield can be calculated from an energy balance, carbon balance, and electron balance.

For further details, see SI 1.3 and Xiao and VanBriesen [22, 24].

< Table 1 >

### 2.4 Conditions for comparison

In all calculations the chemical was assumed to be the sole source of both energy and carbon. Ammonia was taken as the sole nitrogen source so electrons for the assimilatory reduction of  $\text{NO}_3^-$  was not considered, and the carbon was assumed to be used only for cell synthesis or oxidised to  $\text{CO}_2$ , hence, no other carbonaceous products are formed.

Gibbs energy of reaction is calculated using activities of the reactants and products assumed to be 1 M, except for  $\text{H}^+$  which is assumed to be  $10^{-7}$  M (pH 7). The Gibbs free energy of reaction for non-standard conditions can be calculated as

$$\Delta G_r' = \Delta G_r^{o'} + R T \ln Q = \Delta G_r^{o'} + R T \ln \left( \frac{\prod_{i=1}^n [\text{product}]_i^p}{\prod_{i=1}^n [\text{reactant}]_i^r} \right) \quad (1)$$

186  
187 where  $R$  is the ideal gas constant [ $8.314 \text{ J (K mol)}^{-1}$ ],  $T$  is the absolute temperature [K],  
188  $Q$  is the reaction quotient,  $[product]$  and  $[reactant]$  are the activities of products and  
189 reactants, and  $p$  and  $r$  are their respective stoichiometric coefficients. From the equation  
190 itself it can be seen that when  $Q < 1$  the term is negative and when  $Q > 1$  the term is  
191 positive.

192 For aerobic growth,  $\text{O}_2$  was taken as the terminal electron acceptor. For  
193 anaerobic growth both nitrate-reducing conditions ( $\text{NO}_3^-$  as the terminal electron  
194 acceptor) and sulphate-reducing conditions ( $\text{SO}_4^{2-}$  as the terminal electron acceptor)  
195 were investigated. The balanced half-reactions as reductions and their associated Gibbs  
196 free energy of reaction can be found in Table 2 (refer to Table S1, Supplementary  
197 Information for simple carbon substrates). In this table, the thermodynamic frame of  
198 reference can also be seen:  $\text{SO}_4^{2-}$  is reduced to  $\text{H}_2\text{S}$  and  $\text{HS}^-$ ,  $\text{NO}_3^-$  to  $\text{N}_2$ ,  $\text{O}_2$  to  $\text{H}_2\text{O}$ , C  
199 to  $\text{CO}_2$ . Cl is released as  $\text{HCl}$ , and P as  $\text{PO}_4^{2-}$ . The pH was assumed to be 7 and any  
200 change of Gibbs free energy due to speciation was disregarded.

201 For TEEM2, the number of (putative) oxygenase reactions was explicitly taken  
202 into account. However, for ETTYM this was not done. This means that the results  
203 presented in this study differ from those presented in [24].

204 To assess the accuracy and precision of the model predictions, the relative error  
205 ( $E(\%)$ ) and the mean average absolute error (MAE) were calculated.

206

207 < Table 2 >

208

209 *Sensitivity analysis*

To assess the sensitivity of the different parameters on the predicted growth yield, key input parameters were varied including  $Y_{ATP}$ , the assumption of biological standard state conditions (i.e., chemical activities), Gibbs free energy of formation of the compound, and the default bacterial cell formula (and thus the degree of reduction and cellular carbon content). The sensitivity analysis was done changing one-factor-at-a-time, and the parameter sensitivity was evaluated based on the ratio of change in output to the change in input

$$S_i = \frac{\Delta Y}{\Delta X_i} \quad (2)$$

where  $\Delta Y$  is the change in the observed output,  $\Delta X_i$  is the change in input  $i$ , and  $S_i$  is the sensitivity of  $i$ .

### 2.5 Chemicals of environmental concern; data

Thirty-one pesticides were selected for the present study because they are commercially available and widely applied [35] and have fate data (mineralisation and formation of non-extractable residues NER) available in the EU Pesticide Database [36, 37].

Bacterial growth yields have been experimentally assessed only for very few of the selected compounds. Where such data was available, it was used for comparing the performance of the prediction methods. Moreover, ibuprofen and some polycyclic aromatic hydrocarbons (PAH) were also included as the bacterial growth yield has been experimentally determined for these. In total 40 chemicals were selected.

Due to the scarcity of experimentally determined bacterial growth yields on xenobiotics, the methods were also evaluated using the growth yield determined for simple carbon substrates used in biotechnology. The compounds selected for comparison are based on the review and evaluation made by Xiao and VanBriesen [24].

Information regarding the Gibbs free energies of formation, number of carbon-hydrogen bonds,  $Y_{ATP}$ , number of (putative) oxygenase reactions, and degree of reductance are shown along with the name of the compound and reference of Gibbs energy of formation in Table 3 (for simple carbon substrates refer to Table S2).

< Table 3 >

**2.6 Calculation of biogenic non-extractable residues**

Chemicals labelled with carbon isotopes ( $^{14}\text{C}$  or  $^{13}\text{C}$ ) allowed the flow of carbon to be tracked in the experimental system [38-41]. If the compound provides carbon to anabolism and cell synthesis, the labelled carbon will end up in microbial biomass and finally in the biogenic NER. Biogenic NER is not posing a risk to neither the environment nor human health [8]. When a substrate  $S$  is mineralized, the amount of biomass formed is yield times substrate,  $Y \times S$ , and the evolved  $\text{CO}_2$  is  $(1 - Y) \times S$  [15]. After the growth phase has stopped, the maximum ratio between biomass and  $\text{CO}_2$  and is thus

$$\frac{[X_{\text{biogenic NER}}]}{[\text{CO}_2]} = \frac{YS}{(1-Y)S} \text{ or } [X_{\text{biogenic NER}}] = \frac{Y}{1-Y} [\text{CO}_2] \quad (3)$$

where  $X_{\text{biogenic NER}}$  is the biomass making up the living biogenic NER. After the cessation of the growth phase, the microorganisms start to decay. The dead microorganisms are turned over in the microbial food chain and form new biomass,  $\text{CO}_2$  and soil organic matter (SOM) [30-33]. Then, the ratio between biogenic NER and  $^{13/14}\text{CO}_2$  becomes

$$\frac{[\text{SOM}_{\text{biogenic NER}}]}{[\text{CO}_2]} = \frac{f \times Y}{(1-Y) + (1-f) \times Y} \text{ or } [\text{SOM}_{\text{biogenic NER}}] = \frac{f \times Y}{(1-Y) + (1-f) \times Y} [\text{CO}_2] \quad (4)$$

258

259 where  $SOM_{biogenic\ NER}$  is the non-living biogenic NER,  $f$  is the fraction of decaying  
260 biomass turned over into both living biomass and non-living SOM (0.5, [33]), and  $1-f$  is  
261 the fraction of label released as  $CO_2$ . Eq. (3) can be used to estimate NER formation  
262 during short-term experiments, whereas Eq. (4) holds for long-term experiments. It can  
263 be seen that a high mineralization and  $CO_2$  formation together with a high bacterial  
264 growth yield leads to a high formation of biogenic NER.

### 265 3. Results

#### 266 3.1 Comparison of predicted bacterial growth yields

##### 267 *Pesticides and chemicals of environmental concern*

268 In Table 4, the predicted bacterial growth yields under aerobic conditions are shown and  
269 compared to experimentally determined growth yields. In Table 5, the predicted growth  
270 yields under anaerobic conditions are shown.

271

272 *Aerobic conditions.* With oxygen as the terminal electron acceptor, only chlorothalonil  
273 was predicted to have a bacterial growth yield of zero and only by the MTB method.  
274 The reason being the absence of carbon-hydrogen bonds. All other compounds except  
275 pyrene were predicted to have a bacterial growth yield of  $>0.3$  g cell carbon (g substrate  
276 carbon) $^{-1}$  by all methods. The yield predictions from ETTYM were higher than for  
277 TEEM2 which in turn were higher than the predictions by MTB (except for NTA,  
278 where TEEM2 adjusts for energy lost due to one oxygenase reaction). Experimental  
279 yields were found for 13 of the 40 compounds selected. The mean absolute error was  
280 found to be 49% for MTB, 82% for TEEM2, and 111% for ETTYM (Table 4).  
281 Moreover, a strong positive and highly significant ( $p < 0.01$ ) linear correlation between  
282 degree of reductance and predicted yield was found for ETTYM ( $Y = 0.093 \gamma_s + 0.14$ ,

283  $R^2 = 0.70$ ). A weaker but still highly significant ( $p < 0.01$ ) correlation was found for  
 284 TEEM2 ( $Y = 0.09 \gamma_s + 0.15$ ,  $R^2 = 0.53$ ), and a significant ( $p < 0.05$ ) but rather weak  
 285 correlation was found for MTB ( $R^2 = 0.14$ ).

286 For 2,4-D and carbofuran, MTB predicted the yields with an absolute error of  
 287 4.7% and 1.4%, respectively, whereas TEEM2 overestimated them by 51% and 16%,  
 288 respectively, and ETTYM overestimated by 56% and 25%, respectively. For glyphosate  
 289 and anthracene, the predicted growth yields by all three methods deviate greatly from  
 290 the observed value (more than 180% overestimation).

291

292 *Anaerobic conditions.* With nitrate as the electron acceptor, the predicted yields  
 293 decrease 2-5% for all methods (Table 5). All 40 chemicals are predicted to be  
 294 degradable under nitrate-reducing conditions when considering their energy and carbon  
 295 content (the only exception still being chlorothalonil for MTB). The linear correlation  
 296 between yield and degree of reduction was similar to the aerobic case. Again, ETTYM  
 297 predictions were higher than TEEM2 predictions, which in turn are higher than the  
 298 MTB predictions.

299 With sulphate as the terminal electron acceptor, the predicted yields decrease  
 300 54-93% for MTB, 50-118% for TEEM2, and 47-91% for ETTYM, compared to the  
 301 yields found under aerobic conditions (Table 5). For both aerobic and nitrate-reducing  
 302 conditions, the ranking of the estimated yields was close. Under sulphate-reducing  
 303 conditions this was not the case. In fact, TEEM2 predicted negative yields for benzene,  
 304 benzoate, and EDTA (highlighted in bold), while they were positive (albeit low) for  
 305 ETTYM and MTB. MTB predicted 20 chemicals to have a yield of  $<0.1$  g cell carbon  
 306 (g substrate carbon)<sup>-1</sup>, while ten were predicted using TEEM2, and only four using

ETTYM. No linear correlation between the predicted yield and the degree of reductance could be found for any of the methods.

The lack of experimental observations under a multitude of redox conditions makes assessment of the prediction accuracy unachievable.

311

### *Simple carbon substrates*

More experimental data are available for simple carbon substrates, and the result of the growth yield predictions are shown in Table 6. The lowest MAE between experimental and predicted growth yield was observed for TEEM2 (MAE = 15%). MTB showed the highest MAE of 23%. It should be noted, that while TEEM2 took energy losses due to oxygenase reactions into account, the implemented ETTYM method did not.

Subsequently, the work presented in [24] resulted in an even lower MAE when both the pH and related speciation, and oxygenase reactions were taken into account.

All three methods predict oxalate to have the lowest yield. Additionally, the TEEM2 and ETTYM predictions showed similar ranking of the chemicals.

Generally, the highest deviations from the experimental yields were similar for all methods. For MTB, these were formate, oxalate, and glycine; for TEEM2, these were oxalate, glycine, and formate; and for ETTYM, these were oxalate, phenol, and formate.

Excluding oxalate, formate, and glycine from the results, all three methods predicted the experimental growth yield with a MAE below 15%, with TEEM2 still giving the lowest MAE (9%). These three compounds all have a degree of reductance below that of the assumed cell formula for either method.

330

331

< Table 4 >



332

333

< Table 5 >

334

335

< Table 6 >

336

337 **3.2 Sensitivity Analysis**

338 The calculated average sensitivity  $\bar{S}_t$  towards the varied parameters can be found in  
339 Table 7.

340

341  $Y_{ATP}$

342 All methods are relatively sensitive to changes in  $Y_{ATP}$ , especially MTB.  $Y_{ATP}$  is an  
343 uncertain parameter because no such values have ever been determined with pesticides  
344 or other xenobiotics as substrate. The chosen default value for xenobiotics of 5 g cell  
345 dw (mol ATP)<sup>-1</sup> ([18], for methanol) used in the MTB method does not lead to large  
346 errors (cf. Table 4).

347

348 *Gibbs free energy of formation ( $\Delta G_f^{\circ'}$ )*

349 If  $\Delta G_f^{\circ'}$  is positive (e.g. benzene) then an increase in this value would lead to an  
350 increase in the predicted yield. Conversely, a negative  $\Delta G_f^{\circ'}$  which is made more  
351 negative leads to a decrease in the predicted yield (e.g., EDTA). All three methods show  
352 a low sensitivity towards changes of the Gibbs energy of the substrate ( $\bar{S}_t < 0.1$ ). While  
353 TEEM2 and ETTYM also both have low  $\bar{S}_t$  values for changes in  $\Delta G_f^{\circ'}$ , the standard  
354 deviation is high. Yield estimates for chemicals with few carbon atoms (and thus low  
355 formation of CO<sub>2</sub>) and high negative  $\Delta G_f^{\circ'}$  (e.g. glyphosate) are especially sensitive to  
356 changes.

357

358 *Standard state conditions*

359 The deviation from standard state conditions was found to have only a very small effect  
360 on the predicted yields. The relation between  $RT\ln Q$  and  $\Delta G_r'$  in Eq. (1) is logarithmic.  
361 Setting  $\Delta G_r^0$  to zero shows that varying  $Q$  from  $10^{-25}$  to  $10^{25}$  leads to a change in Gibbs  
362 free energy of only  $\pm 143 \text{ kJ mol}^{-1}$  (at standard pressure and temperature).

363

364 *Cell formula*

365 While all methods are sensitive to the cell formula used (Table 7), MTB is the method  
366 least affected. The effect of the cell formula in TEEM2 and ETTYM is not only on the  
367 energy costs related to synthesis [24], but also on the conversion to g cell carbon (g  
368 substrate carbon) $^{-1}$  in TEEM2. This is due to the degree of reduction of the cell ( $\gamma_s$ ) used  
369 in converting the units. For MTB, the predicted yield is only affected by changes in the  
370 assumed cellular carbon content ( $\sigma_C$ ). For MTB, a higher carbon content per mass of  
371 cell leads to a higher yield in g cell carbon (g substrate carbon) $^{-1}$  but lower yield in gram  
372 bacteria (gram substrate) $^{-1}$  (Figure S1).

373

374 **3.3 Prediction of biogenic NER formation based on the predicted growth yields**

375 The growth yields predicted with the MTB method were used to estimate the biogenic  
376 NER formation from the  $\text{CO}_2$  produced during degradation experiments (Table 8). The  
377 formation of biogenic NER was predicted to make up a considerable fraction of the  
378 experimentally determined NER for most of the chemicals. Except for one compound  
379 (glyphosate, caused by the production of the metabolite aminomethylphosphonic acid  
380 (AMPA)), the predicted biogenic NER was smaller than the measured total NER. For  
381 daminozide, almost all of the formed NER is suggested to be made up of biogenic NER

(94%). For MCPA and MCPB, approximately 50% of the NER is biogenic. For bifenazate, iprodione, pendimethalin, phenmedipham, and pymetrozine the biogenic NER is suggested to make up less than 10% of the formed NER. This then suggests that for these chemicals, type I and II NER make up the majority of the formed NER.

The experimental period for ibuprofen and 2,4-D (64 days) [38, 39], and glyphosate and metamitron (80 days) [42, 43] was shorter than the experiments reported in [36]. Eq. (3), which calculates living biomass X as biogenic NER, was additionally used to interpret these experiments. In these four studies, the carbon label found in amino acids was reported. For living microbes, about half of the carbon is in proteins. This fraction increases during decay and turnover of microbial biomass because proteins are the most stable fraction of the cells [30, 33, 15]. Except for glyphosate, the measured label in amino acids is within the range of biogenic NER predicted by Eqs. (3) and (4), and the measured total NER is greater.

< Table 8 >

**4. Discussion**

**4.1 Comparison of predicted bacterial growth yields**

For the pesticides and chemicals of environmental concern, the lack of experimental data for the bacterial growth yield under different redox conditions made it difficult to assess the error associated with the predicted growth yields. Only 14 yield values could be used in the comparison, with some of them used earlier in [15]. Although MTB performs better than TEEM2 and ETTYM, the MAE was still found to be ~50%.

The bacterial growth yield estimation methods are all developed to predict the true yield at optimal growth conditions for microorganisms. The observed value is typically a net yield accounting only for the formation of new cell mass and removal of

the parent compound [15, 23, 24, 44]. The difference between the two is that for the observed yield energy and carbon expenditure, due to non-growth purposes, are not captured (e.g. energy spent on maintenance, formation of metabolites or soluble microbial products and extracellular polymeric substances), unless a dynamic model was used for fit. Hence, the observed yield is typically lower than the true yield. Additionally, the prediction methods assume a complete degradation of the compound. If hardly degradable or insoluble metabolites are formed and rendered not bioavailable (as NER I or II), the observed yield will be lower than the predicted true yield. Compounds which have a known toxic effect (e.g. phenolic compounds [45]) can also result in a higher amount of energy being spent on maintenance leading to an observed yield lower than the predicted true yield.

The bacterial yields for anthracene and glyphosate are by all methods overestimated by >100%. Anthracene is readily adsorbed and is scarcely soluble in water [46]. In the experiments with glyphosate [42], the intermediate AMPA accumulated, resulting in an observed yield much that was lower than the predicted yield (Table 4). If these two are removed from the calculations, the MAE is reduced to 20% for MTB, 40% for TEEM2, and 58% for ETTYM.

The presence of other sources of carbon or energy (mixed substrate use) also adds uncertainty to the observed value. Interestingly, in the experiments with 2,4-D and carbofuran [29, 47], great care was taken in the experimental setup to minimise confounding factors due to other carbon sources, and here the MTB predicted yields are very close to the experimentally determined values.

The observed differences might also be attributed to their high hydrophobicity and limited bioavailability [8, 48-50], which means that truly dissolved concentrations are low. Under these conditions, microbes use most of the growth substrate just for

432 maintenance [51]. Despite the explicit consideration of energy losses related to  
433 (putative) oxygenase reactions for PAHs in TEEM2, its errors were higher than for  
434 MTB. Helbling *et al.* [29] successfully matched the predicted bacterial yield with the  
435 measured bacterial yield on carbofuran by taking oxygenase reactions into account as  
436 suggested in [28].

437 Under sulphate-reducing conditions the predicted bacterial yields were much  
438 lower than the predicted yields under aerobic conditions, which can be expected  
439 considering the lower energy associated with the reduction of sulphate in comparison to  
440 oxygen reduction. An interesting observation was that the decrease in yield is similar  
441 across all three methods. This shows that the half-reaction approach using various  
442 electron acceptors used in ETTYM and TEEM2 can also be used with MTB, which has  
443 not been shown before.

444 Under anaerobic conditions, the energy released from the majority of redox  
445 reactions might not be sufficient to fuel bacterial growth. This suggests that there would  
446 not be an evolutionary incentive to develop metabolic pathways for anaerobic  
447 degradation where the chemical is the electron donor (unless the chemical can provide  
448 other macro- or micronutrients, e.g. nitrogen and carbon). For all methods, the predicted  
449 growth yield is so small that there is no relationship between degree of reductance and  
450 bacterial growth yield.

451 For the simple carbon substrates, both TEEM2 and ETTYM perform better than  
452 MTB (lower MAE). However, one has to consider the fact that the efficiency parameter  
453 in the TEEM2 method was calibrated to the data in order to produce yield estimates  
454 close to experimentally determined growth yields [23]; and experimental yields were  
455 converted to g cell carbon (g substrate carbon)<sup>-1</sup> using the cell formula (CH<sub>2</sub>O<sub>0.6</sub>N<sub>0.2</sub>)  
456 proposed for the ETTYM method [24].

While ETTYM and TEEM2 both overestimated the yield for oxalate, MTB estimated it as zero due to the absence of C-H bonds (which points to the need for a model modification as bacteria are able to grow on this substrate).

#### 4.2 Sensitivity analysis

All the methods were shown to be sensitive to the choice of cell formula but exhibited low sensitivity to variations of the formation energy,  $\Delta G_f^\circ$ , of the chemical of interest. All methods are based on the Gibbs energy of reaction and knowledge of the Gibbs energy of formation of the chemical of interest is needed. If the value has not been determined experimentally (e.g. [52]), it can be estimated using group contribution methods [53-58] (method [53] is implemented in the freely available ChemProp [59]), or by component contribution methods [60] (implemented in the free accessible database eQuilibrator [61]), or calculated using quantum mechanics [62]. For xenobiotics, the applicability of these estimation methods may be limited. Consequently, we also tested the sensitivity of Gibbs energy of formation of the xenobiotic substrate by setting this value to 0 kJ mol<sup>-1</sup> (Table S3). The MTB method has surprisingly low sensitivity. Compounds having a large negative Gibbs energy of formation (e.g. NTA, EDTA, and glyphosate) and few carbon-hydrogen bonds (6, 8, and 4, respectively) show a maximum deviation of around 20% from the predictions done with correct Gibbs energy of formation. Overall, the average deviation is only 6%. In comparison, TEEM2 and ETTYM have considerably higher average deviation (14% and 11%, respectively).

While deviation from the standard state conditions might be needed to render a reaction step thermodynamically feasible [16], the effect on the overall yield prediction would only be seen for reactions where the Gibbs free energy of reaction is low, either

482 due to the low energy associated with the oxidation of substrate (e.g. formate or  
483 formaldehyde), or the low energy associated with the reduction of the electron acceptor  
484 (e.g.  $\text{SO}_4^{2-}$ ). This means that the true concentrations or activities during the reaction can  
485 be neglected without significant error.

486 The effect of pH on Gibbs energy of reaction was not investigated in this study  
487 as this was recently done in [24]. At pH 7, taking the distribution of the inorganic  
488 carbon species into account only changed the predicted yield approximately 1%.  
489 However, speciation of the substrate also has an effect on its Gibbs energy of formation.  
490 Similarly, the sensitivity of the energy capture efficiency parameter  $\varepsilon$  (or variations  
491 thereof) has been assessed elsewhere [22, 23].

493 ***4.3 Prediction of biogenic NER formation and implication for degradation in the***  
494 ***environment***

495 The experiments cited from [36] in Table 8 were run for more than 100 days. The peak  
496 in living biomass is usually after a few days to weeks [38, 39], and therefore we expect  
497 that Eq. (4) ( $\text{SOM}_{\text{biogenic NER}}$ ) is more appropriate for these experiments than Eq. (3)  
498 ( $X_{\text{biogenic NER}}$ ) as the majority of the living biomass has decayed and been incorporated  
499 into SOM after 100 days. Results obtained by Eq. (4) are smaller than the measured  
500 NER, which confirms the results of this equation. The only exception is glyphosate. For  
501 daminozide, the chemical with the highest predicted yield, the calculated biogenic NER  
502 and measured NER are almost equal.

503 The examination of Table 8 gives no significant correlation between measured  
504 total NER and predicted  $X_{\text{biogenic NER}}$  or  $\text{SOM}_{\text{biogenic NER}}$ . Such a correlation should not be  
505 expected since the processes leading to NER I, II and III are competing. If a pesticide is  
506 not degraded it can undergo aging and irreversible sorption (type I NER) and covalent

binding of the parent compound or its metabolites (type II NER) [8]. The estimation of the various fractions of NER can be rebuilt in a dynamic simulation model. We suggested such a model in [8] and used it successfully for the prediction of the NER formation from 2,4-D and ibuprofen with pre-estimated yield data [15].

The data compiled by [36] give no hints into which form the NER are present, and thus cannot serve to validate the estimation equation. However, it is likely that high true yields are connected to high experimental yields, which stimulate bacterial growth and thus microbial degradation. Of course, the enzymatic pathways to facilitate the degradation and energy exploitation of the molecule also need to be present.

In degradation experiments with metamitron [43], glyphosate [42], ibuprofen [39, 63] and 2,4-D [38, 63], the formation of biogenic NER was investigated by tracking the distribution of stable carbon or nitrogen isotope ( $^{13}\text{C}$  or  $^{15}\text{N}$ ) in  $\text{CO}_2$ , amino acids, fatty acids, metabolites, and parent compounds. Experiments of this kind are very helpful to discriminate between the various types of NER and to validate our biogenic NER estimation approach. Shrestha *et al.* [11] observed that the formation of NER occurred simultaneously with the degradation and release of  $\text{CO}_2$ . This shows the coupling of the formation of NER to microbial activity, and to the growth and decay of biomass. Mamy *et al.* [13] observed a lack of QSAR approaches to predict the formation of NER. The method applied in this study provides process-based theoretical background that may be used to interpret NER data derived in degradation experiments. Before routine application though, further confirmation by targeted experiments is still needed.



530 **5. Conclusions**

531 The MTB method was compared with two widely used bacterial growth yield  
532 estimation methods, TEEM2 and ETTYM. The results showed that TEEM2 and  
533 ETTYM methods performed better than MTB in estimating the yield on simple  
534 substrates, while MTB performed better when estimating the yield on organic chemicals  
535 of environmental concern in general and in particular on pesticides. It was also shown  
536 that the MTB approach can be expanded to electron acceptors other than oxygen, like  
537 sulphate and nitrate.

538 The sensitivity analysis revealed that all three methods are relatively sensitive to  
539 changes in  $Y_{ATP}$ , an uncertain parameter. TEEM2 and ETTYM are also sensitive to  
540 changes in the cell formula due to the change in the degree of reductance. All methods  
541 showed low sensitivity to variations in the Gibbs energy of formation of the organic  
542 chemicals because most of the Gibbs energy of reaction stems from the formation of the  
543 oxidation products carbon dioxide and water. The growth yield estimates were then  
544 successfully used to estimate the formation of biogenic non-extractable residues. The  
545 approach applied in this study provides a theoretical foundation that can be used to  
546 predict biogenic NER formation during pesticide degradation experiments. It can also  
547 be employed to interpret NER data derived during regulatory studies.

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556 The MTB theoretical yield tool is available both in spreadsheet and Python code  
557 on request from the first author.

558 **Disclosure statement**

559 The authors declare no financial interest.

560

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561 **References**

562 [1] EU. Regulation (EC) No 1907/2006 of the European Parliament and the Council of  
563 18 December 2006 concerning Registration, Evaluation, Authorisation and  
564 Restriction of Chemicals (REACH), Official Journal of the European Union  
565 (2006), L 136.  
566  
567 [2] EU. Regulation (EC) No 1107/2009 of the European Parliament and the council of  
568 21 October 2009 concerning the placing of plant protection products on the  
569 market and repealing Council Directives 79/117/EEC and 91/414/EEC, Official  
570 Journal of the European Union (2009), L 309/1.  
571  
572 [3] EU. Commission regulation (EU) No 283/2013 setting out the data requirements for  
573 active substances, in accordance with Regulation (EC) No 1107/2009 of the  
574 European Parliament and the Council concerning the placing of plant protection  
575 products on the market. Official Journal of the European Union (2013), L 93/1.  
576  
577 [4] OECD. *Test No. 306: Biodegradability in Seawater*. OECD Publishing, Paris, 1992.  
DOI 10.1787/9789264070486-en  
578  
579 [5] OECD. *Test No. 307: Aerobic and Anaerobic Transformation in Soil*. OECD  
Publishing, Paris, 2002a. DOI 10.1787/9789264070509-en  
580  
581 [6] OECD. *Test No. 308: Aerobic and Anaerobic Transformation in Aquatic Sediment  
Systems*. OECD Publishing, Paris, 2002b. DOI 10.1787/9789264070523-en  
582  
583 [7] OECD. *Test No. 309: Aerobic Mineralisation in Surface Water – Simulation  
Biodegradation Test*. OECD Publishing, Paris, 2004. DOI  
584 10.1787/9789264070547-en  
585  
586 [8] M. Kästner, K. M. Nowak, A. Miltne, S. Trapp, and A. Schäffer, *Classification and  
Modelling of Nonextractable Residue (NER) Formation of Xenobiotics in Soil – A  
587 Synthesis*, Crit. Rev. Env. Sci. Technol. 44 (2014), pp. 2107–2171.  
588  
589 [9] K. Fenner, S. Canonica, L.P. Wackett, and M. Elsner, M., *Evaluating Pesticide  
Degradation in the Environment: Blind Spots and Emerging Opportunities*,  
590 Science 341 (2013), pp. 752–758.  
591  
592 [10] M. Honti, and K. Fenner, *Deriving persistence indicators from regulatory water-  
sediment studies - Opportunities and limitations in OECD 308 data*, Environ. Sci.  
593 Technol. 49 (2015), pp. 5879–5886  
594  
595 [11] P. Shrestha, T. Junker, K. Fenner, S. Hahn, M. Honti, R. Bakkour, C. Diaz, and D.  
Hennecke, *Simulation Studies to Explore Biodegradation in Water–Sediment  
596 Systems: From OECD 308 to OECD 309*, Environ. Sci. Technol. 50 (2016), pp.  
597 6856–6864.  
598  
599 [12] M. Honti, S. Hahn, D. Hennecke, T. Junker, P. Shrestha, and K. Fenner, *Bridging  
across OECD 308 and 309 Data in Search of a Robust Biotransformation  
600 Indicator*, Environ. Sci. Technol. 50 (2016), pp. 6865–6872.

- [13] L. Mamy, D. Patureau, E. Barriuso, C. Bedos, F. Bessac, X. Louchart, F. Martin-Laurent, C. Miege, and P. Benoit, *Prediction of the Fate of Organic Compounds in the Environment From Their Molecular Properties: A Review*, Crit. Rev. Env. Sci. Tec. 45 (2015), pp. 1277–1377.
- [14] A. Sabljic and Y. Nakagawa, Biodegradation and quantitative structure-activity relationship (QSAR), in *Non-First Order Degradation and Time-Dependent Sorption of Organic Chemicals in Soil*, W. Chen, A. Sabljic, S.A. Cryer and R.S. Kookana, eds, Book Series: ACS Symposium Series, American Chemical Society, Washington (DC), Volume 1174, 2014, pp. 57-84.
- [15] S. Trapp, A. Libonati Brock, K. Nowak and M. Kästner, *Prediction of the formation of biogenic non-extractable residues during degradation of environmental chemicals from biomass yields*, Environ. Sci. Technol. (2017), Submitted for publication
- [16] T. Maskow, and U. von Stockar, *How reliable are thermodynamic feasibility statements of biochemical pathways?*, Biotechnol. Bioeng. 92 (2005), pp. 223–230.
- [17] P.L. McCarty, *Thermodynamics of biological synthesis and growth*, Intl. J. Air Water Poll. 9 (1965), pp. 621-639.
- [18] G. Diekert, Grundmechanismen des Stoffwechsels und der Energiegewinnung, in *Umweltbiotechnologie*, J.C.G. Ottow and W. Bidlingmaier, eds., Fischer Verlag, Stuttgart, Germany, 1997; pp. 1-38.
- [19] M.T. Madigan, J. Martinko, and J. Parker, *Biology of Microorganisms*. International Student Edition, Pearson Inc, 2011.
- [20] J.J. Heijnen, *A new thermodynamically based correlation of chemotrophic biomass yields*, Anton. Leeuw. Int. J. G. 60 (1991), pp. 235-256.
- [21] J.J. Heijnen, and J.P. Dijken, *In search of a thermodynamic description of biomass yields for the chemotrophic growth of microorganisms*, Biotechnol Bioeng 39 (1992), pp. 833-852
- [22] J. Xiao, and J.M. VanBriesen, *Expanded thermodynamic model for microbial true yield prediction*, Biotechnol. Bioeng 93 (2006), pp. 110–121.
- [23] P.L. McCarty, *Thermodynamic Electron Equivalents Model for bacterial yield prediction: Modifications and comparative evaluations*, Biotechnol. Bioeng 97 (2007), pp. 377–388.
- [24] J. Xiao, and J.M. VanBriesen, *Expanded thermodynamic true yield prediction model: adjustments and limitations*, Biodegradation 19 (2008), pp. 99–127.
- [25] R. Kleerebezem, and M.C.M. Van Loosdrecht, *A Generalized Method for Thermodynamic State Analysis of Environmental Systems*, Crit. Rev. Env. Sci. Tec. 40 (2010), pp. 1–54.

[26] G. Tchobanoglous, F.L. Burton, and H.D. Stensel, *Wastewater Engineering, Treatment and Reuse*, 4th ed. McGraw-Hill: Boston, USA, 2004.

[27] J.M. VanBriesen, *Evaluation of methods to predict bacterial yield using thermodynamics*, Biodegradation 13 (2002), pp. 171–190.

[28] J.M. VanBriesen, *Thermodynamic yield predictions for biodegradation through oxygenase activation reactions*, Biodegradation 12 (2001), pp. 265–281.

[29] D.E. Helbling, F. Hammes, T. Egli, and H.-P.E. Kohler, *Kinetics and yields of pesticide biodegradation at low substrate concentrations and under conditions restricting assimilable organic carbon*, Appl. Environ. Microbiol. 80 (2014), pp. 1306–1313.

[30] R. Kindler, A. Miltner, H.H. Richnow, and M. Kästner, *Fate of gram-negative bacterial biomass in soil - Mineralization and contribution to SOM*, Soil Biol. Biochem. 38 (2006), pp. 2860–2870.

[31] R. Kindler, A. Miltner, M. Thullner, H.H. Richnow, and M. Kästner, *Fate of bacterial biomass derived fatty acids in soil and their contribution to soil organic matter*, Org. Geochem. 40 (2009), pp. 29–37.

[32] A. Miltner, R. Kindler, H. Knicker, H.H. Richnow, and M. Kästner, *Fate of microbial biomass-derived amino acids in soil and their contribution to soil organic matter*. Org. Geochem. 40 (2009), pp. 978–985.

[33] A. Miltner, P. Bombach, B. Schmidt-Brücken, and M. Kästner, *SOM genesis: Microbial biomass as a significant source*, Biogeochemistry 111 (2012), pp. 41–55.

[34] B.E. Rittmann, P.L. McCarty, *Environmental biotechnology: principles and applications*. McGraw-Hill: New York, NY, 2001.

[35] J. E. Ørum, and H. Hossy, *Bekæmpelsesmiddel-statistik 2014. Behandlingshyppighed og pesticidbelastning, baseret på salgsstatistik og sprøjtejournaldata*. Orientering fra Miljøstyrelsen nr. 13, 2015. Miljøstyrelsen (Danish EPA), 2015.

[36] E. Barriuso, P. Benoit, and I. G. Dubus, *Formation of pesticide nonextractable (bound) residues in soil: Magnitude, controlling factors and reversibility*, Environ. Sci. Technol. 42 (2008), pp. 1845–1854.

[37] EU. *EU Pesticides Database*. Available at <http://ec.europa.eu/food/plant/pesticides/eu-pesticides-database/public/> (visited 2017-04-07).

[38] K. M. Nowak, A. Miltner, M. Gehre, A. Schäffer, and M. Kästner, *Formation and fate of bound residues from microbial biomass during 2, 4-D degradation in soil*, Environ. Sci. Technol. 45 (2011), pp. 999–1006.

- [39] K. M. Nowak, C. Girardi, A. Miltner, M. Gehre, A. Schäffer, and M. Kästner, *Contribution of microorganisms to non-extractable residue formation during biodegradation of ibuprofen in soil*, *Sci. Total Environ.* 445–446 (2013), pp. 377–384.
- [40] C. Poßberg, B. Schmidt, K.M. Nowak, M. Telscher, A. Lagojda, and A. Schaeffer, *Quantitative Identification of Biogenic Nonextractable Pesticide Residues in Soil by <sup>14</sup>C-Analysis* *Environ. Sci. Technol.* 50 (2016), pp. 6415–6422.
- [41] M. Kästner, K.M. Nowak, A. Miltner, and A. Schäffer, *(Multiple) Isotope probing approaches to trace the fate of environmental chemicals and the formation of non-extractable “bound” residues*. *Curr. Opin. Biotech* 41 (2016), pp. 73–82
- [42] S. Wang, B. Seiwert, M. Kästner, A. Miltner, A. Schäffer, T. Reemtsma, Q. Yang, and K.M. Nowak, *(Bio)degradation of glyphosate in water-sediment microcosms - A stable isotope co-labeling approach*, *Water Res* 99 (2016), pp. 91–100.
- [43] S. Wang, A. Miltner, and K.M. Nowak, *Identification of degradation routes of metatriton in soil microcosms using <sup>13</sup>C-isotope labeling*, *Environ. Pollut.* 220 (2017), pp. 927–935.
- [44] D. W. Tempest and O.M. Neijssel, *The Status of  $Y_{app}$  and Maintenance Energy as Biologically Interpretable Phenomena*. *Ann. Rev. Microbiol.*, 38 (1984), pp. 459–486.
- [45] B.I. Escher, R.W. Hunziker, and R.P. Schwarzenbach, *Interaction of phenolic uncouplers in binary mixtures: concentration-additive and synergistic effects*. *Environ Sci Technol* 35 (2001), pp. 3905–14.
- [46] R.P. Schwarzenbach, P.M Gschwend, and D.M. Imboden, *Environmental Organic Chemistry*, 1st ed., John Wiley & Sons, Inc., New York, NY, 1993.
- [47] N.M. Chong, S.C. Tsai, and T.N. Le, *The biomass yielding process of xenobiotic degradation*, *Bioresource Technol.* 101 (2010), pp. 4337–4342.
- [48] T.N. Bosma, P.J.M. Middeldorp, G. Schraa, and A.J.B. Zehnder, *Mass transfer limitation of biotransformation: Quantifying bioavailability*, *Environ. Sci. Technol.* 31 (1997), pp. 248–252.
- [49] L.Y. Wick, T. Colangelo, and H. Harms, *Kinetics of mass transfer-limited bacterial growth on solid PAHs*, *Environ. Sci. Technol.* 35 (2001), pp. 354–361.
- [50] A. Katayama, R. Bhula, G.R. Burns, E. Carazo, A. Felsot, D. Hamilton, C. Harris, Y.H. Kim, G. Kleter, W. Koerdel, J. Linders, J.G.M.W. Peijnenburg, A. Sabljic, R.G. Stephenson, D.K. Racke, B. Rubin, K. Tanaka, J. Unsworth and R.D. Wauchope, *Bioavailability of xenobiotics in the soil environment*, *Rev. Environ. Contam. Toxicol* 203 (2010), pp. 1–86.

- [51] A. Rein, I.K.U. Adam, A. Miltner, K. Brumme, M. Kästner, and S. Trapp, *Impact of bacterial activity on turnover of insoluble hydrophobic substrates (phenanthrene and pyrene) — Model simulations for prediction of bioremediation success*, J. Hazard. Mater. 306 (2016), pp. 105–114.
- [52] R.K. Thauer, K. Jungermann, and K. Decker, *Energy conservation in chemotrophic anaerobic bacteria*, Bacteriol. Rev. 41 (1977), pp. 100–180.
- [53] K. G. Joback, and R.C. Reid, *Estimation of pure-component properties from group-contributions*, Chem. Eng. Comm. 57 (1987), pp. 233–243.
- [54] M.L. Mavrovouniotis, *Group contributions for estimating standard Gibbs energies of formation of biochemical compounds in aqueous solution*, Biotech. Bioeng. 36 (1990), pp. 1070–1082.
- [55] M. L. Mavrovouniotis, *Estimation of standard Gibbs energy changes of biotransformations*, J. Biol. Chem. 266 (1991), pp. 14440–14445.
- [56] M.L. Mavrovouniotis, S. Prickett, and L. Constantinou, *Object-oriented estimation of properties from molecular structure*, Comput. Chem. Eng. 16 (1992), pp. 353–360.
- [57] J. Marrero, and R. Gani, *Group-contribution based estimation of pure component properties*, Fluid Phase Equilibr. 183–184 (2001), pp. 183–208.
- [58] A. S. Hukkerikar, B. Sarup, A. Ten Kate, J. Abildskov, G. Sin, and R. Gani, *Group-contribution+ (GC+) based estimation of properties of pure components: Improved property estimation and uncertainty analysis*, Fluid Phase Equilibr. 321 (2012), pp. 25–43.
- [59] G. Schüürmann, R. Kühne, F. Kleint, R.-U. Ebert, C. Rothenbacher, and P. Herth, *A software system for automatic chemical property estimation from molecular structure*. in *Quantitative Structure-Activity Relationships in Environmental Sciences – VII*, F. Chen and G. Schüürmann G, eds, SETAC Press, Pensacola (FL), USA, 1996, pp. 93–114.
- [60] E. Noor, H.S. Haraldsdóttir, R. Milo, R.M.T. Fleming, *Consistent Estimation of Gibbs Energy Using Component Contributions*, PLoS Comput. Biol. 9 (2013).
- [61] A. Flamholz, E. Noor, A. Bar-Even, and R. Milo, *eQuilibrator - the biochemical thermodynamics calculator*, Nucleic Acids Res. 40 (2012), pp. 770–775.
- [62] *Mol-Instincts, A Fundamental Chemical Database based on Quantum Mechanics & QSPR*, Copyright by ChemEssen, Inc. 2017; software available at <http://www.mol-in.com>
- [63] C. Girardi, K.M. Nowak, O. Carranza-Diaz, B. Lewkow, A. Miltner, M. Gehre, A. Schäffer, M. Kästner, *Microbial degradation of the pharmaceutical ibuprofen and the herbicide 2,4-D in water and soil - Use and limits of data obtained from*

- 1  
2  
3 753        *aqueous systems for predicting their fate in soil*, Sci. Total Environ. 444 (2013),  
4 754        pp. 32–42.
- 5  
6 755        [64] S.L. Trabue, X. Feng, A.V. Ogram, and L.T. Ou, *Carbofuran degradation in soil*  
7 756        *profiles*, J. Environ. Sci. Heal. B 32 (19976), pp. 861–878.
- 8  
9 757        [65] Z. Yuan and J. M. VanBriesen, *Yield prediction and stoichiometry of multi-step*  
10 758        *biodegradation reactions involving oxygenation*, Biotechnol. Bioeng. 80 (2002),  
11 759        pp. 100–113
- 12  
13  
14 760        [66] C.N Sawyer, P.L. McCarty, and G.F. Parkin, *Chemistry for environmental*  
15 761        *engineering and science*, 5th Edn., McGraw-Hill: New York, USA, 752 p., 2003.
- 16  
17 762        [67] P.W. Atkins, and J. de Paula, *Physical Chemistry for the Life Sciences*. 1st ed.  
18  
19 763        Oxford University Press: Oxford, UK. 2006.
- 20  
21 764        [68] L. Toräng, N. Nyholm, H.-J. Albrechtsen, *Shifts in biodegradation kinetics of the*  
22 765        *herbicides MCPP and 2,4-D at low concentrations in aerobic aquifer materials*,  
23 766        Environ. Sci. Technol 37 (2003), pp. 3095–3103.
- 24  
25 767        [69] Z. Yuan, *The role of oxygenation and intermediates in biodegradation of chelating*  
26 768        *agents*, Ph.D. diss., Carnegie Mellon University, US, 2004.
- 27  
28 769        [70] J.D. Linton and R.J. Stephenson, *A preliminary study on growth yields in relation*  
29 770        *to the carbon and energy content of various organic growth substrates*, FEMS  
30 771        Microbiol. Lett. 3 (1978), pp. 95–98  
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773 **Tables:**

774 Table 1. Comparison of equations used to estimate the bacterial yield on a given substrate  
775 serving as both electron donor and carbon source in g cell carbon (g substrate carbon)<sup>-1</sup>. The  
776 nitrogen source is assumed to be NH<sub>4</sub><sup>+</sup>. Carbon is assumed to be incorporated into biomass or  
777 evolved as CO<sub>2</sub>.

Method	Equation
Minimum Turnover to Biomass (MTB)	$Y_{\frac{C}{C}} = \left( \frac{\frac{n_{bio}}{n} \frac{\Delta G_r^{0'}}{\Delta G_{ATP}} \times Y_{ATP}}{\frac{M_c}{\sigma_c} \times n_c + \frac{n_{bio}}{n} \frac{\Delta G_r^{0'}}{\Delta G_{ATP}} \times Y_{ATP}} \right)$
Thermodynam ic Electron Equivalent Model 2 (TEEM2)	$Y_{\frac{C}{C}} = \left( \frac{\gamma_s}{\gamma_c} \right) \left( \frac{\Delta G_a^{0'} - \Delta G_d^{0'} - \frac{q}{p} \Delta G_{xy}^{0'}}{\Delta G_a^{0'} - \Delta G_d^{0'} - \frac{q}{p} \Delta G_{xy}^{0'} - \frac{\Delta G_{fa}^{0'} - \Delta G_d^{0'}}{\epsilon^m} + \frac{\Delta G_{in}^{0'} - \Delta G_{fa}^{0'}}{\epsilon^n} + \frac{\frac{\Delta G_{ATP}}{Y_{ATP} \times 0.9} \times \frac{M_c}{\gamma_c \sigma_c}}{\epsilon} \right)$
Expanded thermodynami c true yield model (ETTYM)	$Y_{\frac{C}{C}} = f_{cell} = \frac{-K (\gamma_s \Delta G_{e-o_2} - \Delta G_d)}{\frac{\Delta G_{acetate} - \Delta G_d}{K^m} + \frac{\frac{\Delta G_{ATP} \times M_{cell}}{Y_{ATP} \times 0.9}}{K} - K (\gamma_c \Delta G_{e-o_2} - \Delta G_d)}$

778 The parameters are: MTB;  $n_{bio}$ : bio-available electrons;  $n$ : electrons transferred in the redox reaction;  $\Delta G_r^{0'}$ : Gibbs free energy of  
779 the redox reaction;  $\Delta G_{ATP}$ : Gibbs free energy of hydrolysis with ~40% efficiency taken into account [80 kJ mol<sup>-1</sup>];  $Y_{ATP}$ : bacterial  
780 yield on ATP [g cell dw (mol ATP)<sup>-1</sup>], which is assumed to be dependent on the chemical structure;  $M_c$ : molar mass of carbon  
781 [12.01 g mol<sup>-1</sup>];  $\sigma_c$ : fraction of carbon in dry cell [g C (g cell dw)<sup>-1</sup>];  $n_c$ : number of carbon atoms in the substrate.  
782 TEEM2:  $\gamma_s$ : degree of reductance of the substrate;  $\gamma_c$ : degree of reductance of the cell;  $\Delta G_a^{0'}$ : Gibbs free energy of reduction of the  
783 electron acceptor [kJ eeq<sup>-1</sup>];  $\Delta G_d^{0'}$ : Gibbs free energy of reduction of the electron donor [kJ eeq<sup>-1</sup>];  $q$ : number of oxygenase  
784 reactions [oxygenase reactions mol<sup>-1</sup>];  $p$ : number of electron equivalents per mole substrate [eeq mol<sup>-1</sup>];  $\Delta G_{xy}$ : reduction potential of  
785 NADH/NAD<sup>+</sup> oxidation [= -219.2 kJ mol<sup>-1</sup>];  $\Delta G_{fa}^{0'}$ : Gibbs free energy of reduction of formaldehyde [= 46.53 kJ eeq<sup>-1</sup>];  $\Delta G_{in}^{0'}$ :  
786 Gibbs free energy of reduction of acetyl-CoA [= 30.9 kJ eeq<sup>-1</sup>];  $\epsilon$ : energy capture efficiency [=0.37];  $\Delta G_{ATP}$ : hydrolysis of ATP at  
787 standard biological conditions [= 30.53 kJ mol<sup>-1</sup>];  $Y_{ATP}$ : the bacterial yield on ATP [=10.5 g cell dw (mol ATP)<sup>-1</sup>];  $\sigma_c$ : fraction of  
788 carbon in the cell;  $M_c$ : molar mass of carbon [12.01 g mol<sup>-1</sup>];  $m$ : +1 if  $\Delta G_{fa}^{0'} > 0$ , else =  $n$ ;  $n$ : +1 if  $\Delta G_{in}^{0'} - \Delta G_d^{0'} > 0$ , else  $n = -1$ .

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3 789 ETTYM:  $K$ : efficiency parameter [=0.41];  $\gamma_s$ : degree of reductance of the substrate;  $\Delta G_{e-O_2}$ : Gibbs free energy of reduction of  
4 790 oxygen [ $\text{kJ eq}^{-1}$ ];  $\gamma_c$ : degree of reductance of the cell;  $\Delta G_d$ : Gibbs free energy of the carbon source [ $\text{kJ (mol C)}^{-1}$ ];  $\Delta G_{acetate}$ : Gibbs  
5 791 free energy of acetate reduction ( $= 106.3 \text{ kJ (mol C)}^{-1}$ );  $\Delta G_{ATP}$ : hydrolysis of ATP at standard biological conditions [ $= 30.53 \text{ kJ mol}^{-1}$ ];  
6 792  $Y_{ATP}$ : bacterial yield on ATP [ $= 10.5 \text{ g cell dw (mol ATP)}^{-1}$ ];  $M_{cell}$ : cell mass per mol carbon ( $= 26.4 \text{ g (mol C)}^{-1}$ ) with cell formula  
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8 793  $\text{C}_5\text{H}_{10}\text{O}_3\text{N}$ );  $m$ :  $m$  is +1 if  $(\Delta G_{acetate} - \Delta G_{CS}) > 0$  else  $m$  is = -1.  
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Table 2. Balanced half-reactions as reductions of the three different terminal electron acceptors and pesticides and xenobiotics and their associated Gibbs free energy of the half-reaction ( $\Delta G_r^\circ$ ) in  $\text{kJ mol}^{-1}$  and  $\text{kJ (electron equivalent (eeq))}^{-1}$  at standard state conditions, except for  $\text{H}^+$  ( $=10^{-7} \text{ M}$ ).

Terminal electron acceptor	Half-reaction	$\Delta G_r^\circ$	
		[ $\text{kJ mol}^{-1}$ ]	[ $\text{kJ eeq}^{-1}$ ]
Oxygen, $\text{O}_2$	$\text{O}_2 + 4\text{H}^+ + 4\text{e}^- \rightleftharpoons 2 \text{H}_2\text{O}$	-314.88	-78.72
Nitrate, $\text{NO}_3^-$	$\text{NO}_3^- + 6\text{H}^+ + 5\text{e}^- \rightleftharpoons \frac{1}{2} \text{N}_2 + 3 \text{H}_2\text{O}$	-358.8	-71.76
Sulphate, $\text{SO}_4^{2-}$	$\text{SO}_4^{2-} + 9.5\text{H}^+ + 8\text{e}^- \rightleftharpoons \frac{1}{2} \text{HS}^- + \frac{1}{2} \text{H}_2\text{S} + 4 \text{H}_2\text{O}$	170.16	21.27
Electron donor, Pesticides and xenobiotics			
2,4-D	$8 \text{CO}_2 + 2 \text{HCl} + 30 \text{H}^+ + 30 \text{e}^- \rightleftharpoons \text{C}_8\text{H}_6\text{Cl}_2\text{O}_3 + 13 \text{H}_2\text{O}$	1286	42.9
2,4-DB	$10 \text{CO}_2 + 2\text{HCl} + 42 \text{H}^+ + 42 \text{e}^- \rightleftharpoons \text{C}_{10}\text{H}_{10}\text{Cl}_2\text{O}_3 + 17 \text{H}_2\text{O}$	1778	42.3
Acetamiprid	$10 \text{CO}_2 + \text{HCl} + 4\text{NH}_3 + 38\text{H}^+ + 38\text{e}^- \rightleftharpoons \text{C}_{10}\text{H}_{11}\text{ClN}_4 + 20\text{H}_2\text{O}$	1695	44.6
Acetochlor	$14 \text{CO}_2 + \text{HCl} + \text{NH}_3 + 68\text{H}^+ + 68\text{e}^- \rightleftharpoons \text{C}_{14}\text{H}_{20}\text{ClNO}_2 + 26\text{H}_2\text{O}$	2091	30.8
Alachlor	$14 \text{CO}_2 + 68 \text{H}^+ + 68 \text{e}^- + \text{HCl} +$	2890	42.5

	$\text{NH}_3 \rightleftharpoons \text{C}_{14}\text{H}_{20}\text{ClNO}_2 + 26 \text{H}_2\text{O}$		
Anthracene	$14 \text{CO}_2 + 66\text{H}^+ + 66\text{e}^-$ $\rightleftharpoons \text{C}_{14}\text{H}_{10} + 28\text{H}_2\text{O}$	2204	33.4
Atrazine	$8 \text{CO}_2 + \text{HCl} + 5\text{NH}_3 + 30\text{H}^+ + 30\text{e}^-$ $\rightleftharpoons \text{C}_8\text{H}_{14}\text{ClN}_5$ $+ 16\text{H}_2\text{O}$	1629	54.3
Azoxystrobin	$22 \text{CO}_2 + 3\text{NH}_3 + 86\text{H}^+ + 86\text{e}^-$ $\rightleftharpoons \text{C}_{22}\text{H}_{17}\text{N}_3\text{O}_5$ $+ 39\text{H}_2\text{O}$	3406	39.6
Benalxylyl	$20 \text{CO}_2 + 94 \text{H}^+ + 94 \text{e}^- + \text{NH}_3$ $\rightleftharpoons \text{C}_{20}\text{H}_{23}\text{NO}_3$ $+ 37 \text{H}_2\text{O}$	3652	38.9
Benzene	$6 \text{CO}_2 + 30 \text{H}^+ + 30 \text{e}^-$ $\rightleftharpoons \text{C}_6\text{H}_6 + 12 \text{H}_2\text{O}$	848	28.3
Benzoate	$7 \text{CO}_2 + 31 \text{H}^+ + 32 \text{e}^-$ $\rightleftharpoons \text{C}_7\text{H}_7\text{O}_2^- + 12 \text{H}_2\text{O}$	1043	32.6
Bifenazate	$17 \text{CO}_2 + 76 \text{H}^+ + 76 \text{e}^- + 2 \text{NH}_3$ $\rightleftharpoons \text{C}_{17}\text{H}_{20}\text{N}_2\text{O}_3$ $+ 31 \text{H}_2\text{O}$	3395	44.7
Carbofuran	$12 \text{CO}_2 + \text{NH}_3 + 54\text{H}^+ + 54\text{e}^-$ $\rightleftharpoons \text{C}_{12}\text{H}_{15}\text{NO}_3$ $+ 21\text{H}_2\text{O}$	1676	31.0
Chlorothalonil	$8 \text{CO}_2 + 22 \text{H}^+ + 22 \text{e}^- + 4 \text{HCl}$ $+ 2 \text{NH}_3$ $\rightleftharpoons \text{C}_8\text{Cl}_4\text{N}_2 + 16 \text{H}_2\text{O}$	977	44.4

Chlorpropham	$10 \text{ CO}_2 + 44 \text{ H}^+ + 44 \text{ e}^- + \text{HCl} + \text{NH}_3$ $\rightleftharpoons \text{C}_{10}\text{H}_{12}\text{ClNO}_2$ $+ 18 \text{ H}_2\text{O}$	2223	50.5
Cypermethrin	$22 \text{ CO}_2 + 96 \text{ H}^+ + 96 \text{ e}^- + 2 \text{ HCl}$ $+ \text{NH}_3$ $\rightleftharpoons \text{C}_{22}\text{H}_{19}\text{Cl}_2\text{NO}_3$ $+ 41 \text{ H}_2\text{O}$	3762	39.2
Daminozide	$6 \text{ CO}_2 + 2 \text{ NH}_3 + 24 \text{ H}^+ + 24 \text{ e}^-$ $\rightleftharpoons \text{C}_6\text{H}_{12}\text{N}_2\text{O}_3$ $+ 9 \text{ H}_2\text{O}$	1251	52.1
DDT	$14 \text{ CO}_2 + 60 \text{ H}^+ + 60 \text{ e}^- + 5 \text{ HCl}$ $\rightleftharpoons \text{C}_{14}\text{H}_9\text{Cl}_5 + 28 \text{ H}_2\text{O}$	2508	41.8
Desmedipham	$16 \text{ CO}_2 + 2 \text{ NH}_3 + 66 \text{ H}^+ + 66 \text{ e}^-$ $\rightleftharpoons \text{C}_{16}\text{H}_{16}\text{N}_2\text{O}_4$ $+ 28 \text{ H}_2\text{O}$	3342	50.6
Dicamba	$8 \text{ CO}_2 + 30 \text{ H}^+ + 30 \text{ e}^- + 2 \text{ HCl} \rightleftharpoons$ $\text{C}_8\text{H}_6\text{Cl}_2\text{O}_3 + 13 \text{ H}_2\text{O}$	1278	42.6
Ethylenediaminetetraacetate (EDTA)	$10 \text{ CO}_2 + 2 \text{ NH}_3 + 34 \text{ H}^+ + 34 \text{ e}^-$ $\rightleftharpoons \text{C}_{10}\text{H}_{16}\text{N}_2\text{O}_8$ $+ 12 \text{ H}_2\text{O}$	1294	38.1
Famoxadone	$22 \text{ CO}_2 + 2 \text{ NH}_3 + 92 \text{ H}^+ + 92 \text{ e}^-$ $\rightleftharpoons \text{C}_{22}\text{H}_{18}\text{N}_2\text{O}_4$ $+ 40 \text{ H}_2\text{O}$	3840	41.7
Glyphosate	$3 \text{ CO}_2 + 15 \text{ H}^+ + 12 \text{ e}^- + \text{PO}_4^{3-} + \text{NH}_3$ $\rightleftharpoons \text{C}_3\text{H}_8\text{NO}_5\text{P} + 5 \text{ H}_2\text{O}$	756	63.0
Ibuprofen	$13 \text{ CO}_2 + 66 \text{ H}^+ + 66 \text{ e}^- +$ $\rightleftharpoons \text{C}_{13}\text{H}_{18}\text{O}_2$ $+ 24 \text{ H}_2\text{O}$	2566	38.9

Iprodione	$13 \text{ CO}_2 + 2\text{HCl} + 3\text{NH}_3 + 48\text{H}^+$ $+ 48\text{e}^-$ $\rightleftharpoons \text{C}_{13}\text{H}_{13}\text{Cl}_2\text{N}_3\text{O}_3$ $+ 23\text{H}_2\text{O}$	2358	49.1
MCPA	$9 \text{ CO}_2 + 38 \text{ H}^+ + 38 \text{ e}^- + \text{HCl}$ $\rightleftharpoons \text{C}_9\text{H}_9\text{ClO}_3$ $+ 15 \text{ H}_2\text{O}$	1530	40.3
MCPB	$11 \text{ CO}_2 + \text{HCl} + 50\text{H}^+ + 50\text{e}^-$ $\rightleftharpoons \text{C}_{11}\text{H}_{13}\text{ClO}_3$ $+ 19\text{H}_2\text{O}$	1964	39.3
Mecoprop (MCP)	$10 \text{ CO}_2 + 44 \text{ H}^+ + 44 \text{ e}^- + \text{HCl}$ $\rightleftharpoons \text{C}_{10}\text{H}_{11}\text{ClO}_3$ $+ 17 \text{ H}_2\text{O}$	1779	40.4
Metalaxyl-M	$15 \text{ CO}_2 + \text{NH}_3 + 70\text{H}^+ + 70\text{e}^-$ $\rightleftharpoons \text{C}_{15}\text{H}_{21}\text{NO}_4$ $+ 26\text{H}_2\text{O}$	2997	42.8
Metamitron	$10 \text{ CO}_2 + 36 \text{ H}^+ + 36 \text{ e}^- + 4 \text{ NH}_3$ $\rightleftharpoons \text{C}_{10}\text{H}_{10}\text{N}_4\text{O}$ $+ 19 \text{ H}_2\text{O}$	1391	38.6
Milbemectin	$31 \text{ CO}_2 + 154\text{H}^+ + 154\text{e}^-$ $\rightleftharpoons \text{C}_{31}\text{H}_{44}\text{O}_7 + 55\text{H}_2\text{O}$	6402	41.6
Naphthalene	$10 \text{ CO}_2 + 48\text{H}^+ + 48\text{e}^-$ $\rightleftharpoons \text{C}_{10}\text{H}_8 + 20\text{H}_2\text{O}$	1638	34.1
Nitrilotriacetate (NTA)	$6\text{CO}_2 + \text{NH}_3 + 18\text{H}^+ + 18\text{e}^-$ $\rightleftharpoons \text{C}_6\text{H}_9\text{NO}_6 + 6\text{H}_2\text{O}$	731	40.6
Paraquat	$12 \text{ CO}_2 + 2\text{NH}_3 + 56\text{H}^+ + 56\text{e}^-$ $\rightleftharpoons \text{C}_{12}\text{H}_{14}\text{N}_2 + 24\text{H}_2\text{O}$	2229	39.8

Pendimethalin	$13 \text{ CO}_2 + 54 \text{ H}^+ + 54 \text{ e}^- + 3 \text{ NH}_3$ $\rightleftharpoons \text{C}_{13}\text{H}_{19}\text{N}_3\text{O}_4$ $+ 22 \text{ H}_2\text{O}$	3082	57.1
Phenanthrene	$14 \text{ CO}_2 + 66 \text{ H}^+ + 66 \text{ e}^-$ $\rightleftharpoons \text{C}_{14}\text{H}_{10} + 28 \text{ H}_2\text{O}$	2204	33.4
Phenmedipham	$16 \text{ CO}_2 + 66 \text{ H}^+ + 66 \text{ e}^- + 2 \text{ NH}_3$ $\rightleftharpoons \text{C}_{16}\text{H}_{16}\text{N}_2\text{O}_4$ $+ 28 \text{ H}_2\text{O}$	3336	50.5
Propyzamid	$12 \text{ CO}_2 + 52 \text{ H}^+ + 52 \text{ e}^- + 2 \text{ HCl}$ $+ \text{NH}_3$ $\rightleftharpoons \text{C}_{12}\text{H}_{11}\text{Cl}_2\text{NO}$ $+ 23 \text{ H}_2\text{O}$	2036	39.2
Pymetrozine	$11 \text{ CO}_2 + 5 \text{ NH}_3 + 38 \text{ H}^+ + 38 \text{ e}^-$ $\rightleftharpoons \text{C}_{11}\text{H}_{11}\text{N}_5\text{O}$ $+ 21 \text{ H}_2\text{O}$	1880	49.5
Pyrene	$16 \text{ CO}_2 + 74 \text{ H}^+ + 74 \text{ e}^-$ $\rightleftharpoons \text{C}_{16}\text{H}_{10} + 32 \text{ H}_2\text{O}$	1968	26.6

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Table 3. Gibbs free energy of formation in  $\text{kJ mol}^{-1}$  ( $\Delta G_f^\circ$ ), number of carbon-hydrogen bonds,  $Y_{ATP}$  in g dry weight ( $\text{mol ATP}^{-1}$ ), number of oxygenase reactions ( $t_{oxy}$ ), chemical structure, and degree of reductance ( $\gamma_s$ ) of the pesticides and xenobiotics used in the comparison.

Name	Structure	$\Delta G_f^\circ$	C-H bonds	$Y_{ATP}$	$t_{oxy}$	$\gamma_s$	Reference
2,4-D	$\text{C}_8\text{H}_6\text{Cl}_2\text{O}_3$	-241.5	5	5	0	3.75	[61]
2,4-DB	$\text{C}_{10}\text{H}_{10}\text{Cl}_2\text{O}_3$	-67.8	9	5	0	4.20	[61]
Acetamiprid	$\text{C}_{10}\text{H}_{11}\text{ClN}_4$	745.6	11	5	0	3.80	[61]
Acetochlor	$\text{C}_{14}\text{H}_{20}\text{ClNO}_2$	-128.1	20	5	0	4.86	[61]
Alachlor	$\text{C}_{14}\text{H}_{20}\text{ClNO}_2$	670.8	20	5	0	4.86	[61]
Anthracene	$\text{C}_{14}\text{H}_{10}$	695.8	10	5	1	4.71	[61]
Atrazine	$\text{C}_8\text{H}_{14}\text{ClN}_5$	811.3	12	5	0	3.75	[61]
Azoxystrobin	$\text{C}_{22}\text{H}_{17}\text{N}_3\text{O}_5$	478.1	17	5	0	3.91	[61]
Benalaxyl	$\text{C}_{20}\text{H}_{23}\text{NO}_3$	771.4	23	5	0	4.70	[61]
Benzene	$\text{C}_6\text{H}_6$	133.9	6	5	2	5.00	[66] cited in [23]
Benzoate	$\text{C}_7\text{H}_7\text{O}_2^-$	-105.4	6	5	2	4.57	[62]
Bifenazate	$\text{C}_{17}\text{H}_{20}\text{N}_2\text{O}_3$	965.2	18	5	0	4.47	[61]
Carbofuran	$\text{C}_{12}\text{H}_{15}\text{NO}_3$	-251.6	14	5	2	4.50	[61]
Chlorothalonil	$\text{C}_8\text{Cl}_4\text{N}_2$	163.8	0	5	0	2.75	[61]
Chlorpropham	$\text{C}_{10}\text{H}_{12}\text{ClNO}_2$	639.1	11	5	0	4.40	[61]
Cypermethrin	$\text{C}_{22}\text{H}_{19}\text{Cl}_2\text{NO}_3$	700	19	5	0	4.36	[61]
Daminozide	$\text{C}_6\text{H}_{12}\text{N}_2\text{O}_3$	11.2	11	5	0	4.00	[61]
DDT	$\text{C}_{14}\text{H}_9\text{Cl}_5$	583	9	5	0	4.29	[61]
Desmedipham	$\text{C}_{16}\text{H}_{16}\text{N}_2\text{O}_4$	993	14	5	0	4.13	[61]
Dicamba	$\text{C}_8\text{H}_6\text{Cl}_2\text{O}_3$	-249.8	5	5	0	3.75	[61]
Ethylenediaminetetraacetate (EDTA)	$\text{C}_{10}\text{H}_{16}\text{N}_2\text{O}_8$	-1209.2	8	5	4	3.40	[65]
Famoxadone	$\text{C}_{22}\text{H}_{18}\text{N}_2\text{O}_4$	935.9	17	5	0	4.18	[61]
Glyphosate	$\text{C}_3\text{H}_8\text{NO}_5\text{P}$	-883.5	4	5	0	4.00	[61]
Iprodione	$\text{C}_{13}\text{H}_{13}\text{Cl}_2\text{N}_3\text{O}_3$	434	12	5	0	5.08	[61]



Ibuprofen	$C_{13}H_{18}O_2$	504	17	5	0	3.69	[61]
MCPA	$C_9H_9ClO_3$	-105.2	8	5	0	4.22	[61]
MCPB	$C_{11}H_{13}ClO_3$	11.2	12	5	0	4.55	[61]
Mecoprop (MCP)	$C_{10}H_{11}ClO_3$	-15.7	10	5	0	4.40	[61]
Metalaxyl-M	$C_{15}H_{21}NO_4$	434.8	21	5	0	4.67	[61]
Metamitron	$C_{10}H_{10}N_4O$	414.8	8	5	0	3.60	[62]
Milbemectin	$C_{31}H_{44}O_7$	1090	42	5	0	4.97	[61]
Naphthalene	$C_{10}H_8$	527.1	10	5	1	4.80	[61]
Nitilotriacetate (NTA)	$C_6H_9NO_6$	-954.8	6	5	1	3.00	[65]
Paraquat	$C_{12}H_{14}N_2$	906.4	14	5	0	4.67	[61]
Pendimethalin	$C_{13}H_{19}N_3O_4$	944.3	18	5	0	4.15	[61]
Phenanthrene	$C_{14}H_{10}$	695.8	10	5	1	4.71	[61]
Phenmedipham	$C_{16}H_{16}N_2O_4$	986.9	14	5	0	4.13	[61]
Propyzamide	$C_{12}H_{11}Cl_2NO$	399.9	10	5	0	4.33	[61]
Pymetrozine	$C_{11}H_{11}N_5O$	878	10	5	0	3.45	[61]
Pyrene	$C_{16}H_{10}$	301.3	10	5	1	4.63	[62]
Ammonia	$NH_3$	-26.6					[52]
Carbon dioxide	$CO_2$	-394.4				0	[52]
Nitrate	$NO_3^-$	-108.7					[67]
Hydrogen ion (proton) (pH 7)	$H^+$	-39.9					[52]
Water	$H_2O$	-237.2					[52]
Oxygen	$O_2$	0					[67]
Hydrogen sulphide anion	$HS^-$	12.1					[52]
Hydrogen sulphide	$H_2S$	-27.8					[67]
Sulphate	$SO_4^{2-}$	-744.4					[52]
Nitrogen	$N_2$	0					[67]
Hydrogen chloride	$HCl$	-131.2					[67]

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Table 4. Bacterial growth yields on organic chemicals of environmental concern under aerobic conditions in g cell carbon (g substrate carbon)<sup>-1</sup> predicted using MTB, TEEM2, and MTB. Predictions are evaluated using available experimental data. The error and mean absolute error are shown. The observed experimental bacterial growth yields ( $Y^{OBS}$ ) with reference are shown. The cell formulation for MTB and TEEM2 was taken to be C<sub>5</sub>H<sub>7</sub>O<sub>2</sub>N with a degree of reductance of 4.0. For ETTYM it was taken to be C<sub>5</sub>H<sub>10</sub>O<sub>3</sub>N with a degree of reductance of 4.2. The predictions were made under standard state conditions (pH = 7). The entries are sorted from low to high predicted yield of MTB. For TEEM2 a weak positive correlation and for ETTYM a strong positive correlation exists between the degree of reductance of the compound and predicted yield ( $R^2$  is shown).

Compound	$Y^{OBS}$	$Y^{pred}$ MTB	Error	$Y^{pred}$ TEEM2	Error	$Y^{pred}$ ETTYM	Error	Reference
Unit	g cell carbon (g substrate carbon) <sup>-1</sup>	g cell carbon (g substrate carbon) <sup>-1</sup>	[%]	g cell carbon (g substrate carbon) <sup>-1</sup>	[%]	g cell carbon (g substrate carbon) <sup>-1</sup>	[%]	
Chlorothalonil		0.00		0.35		0.36		
Pyrene	0.21- 0.31	0.27	-13 to 27	0.44	45 to 111	0.52	145	[51]
2,4-D	0.31	0.30	-4.7	0.47	51	0.48	56	[47]
Dicamba		0.30		0.47		0.48		
DDT		0.30		0.53		0.55		
Anthracene	0.11- 0.13	0.31	128 to 183	0.54	182 to 341	0.56	316 to 416	[49]
Phenanthrene	0.32	0.31	-4.1	0.54	67.3	0.56	75	[23]

Azoxystrobin		0.34		0.48		0.49		
Famoxadone		0.34		0.52		0.53		
Ethylenediaminetetraacetate (EDTA)	0.27	0.34	26	0.36	31.6	0.42	55	[69] cited in [24]
Metamitron	0.30	0.34	14	0.43	45	0.45	49	[46]
Benzoate	0.42	0.35	-18	0.49	16	0.54	28	[70] cited in [24]
Propyzamide		0.35		0.52		0.54		
Cypermethrin		0.36		0.53		0.55		
MCPA		0.37		0.52		0.53		
Benzene	0.43	0.37	-13	0.47	10	0.57	33	[23]
2,4-DB		0.38		0.52		0.54		
Naphthalene	0.47	0.38	-18	0.55	16	0.57	22	[23]
Desmedipham		0.39		0.55		0.57		
Phenmedipham		0.39		0.55		0.56		
Pymetrozine		0.39		0.45		0.47		
Iprodione		0.40		0.49		0.50		
Mecoprop (MCPP)	0.3	0.40	32	0.54	80	0.56	85	[68]
Nitritotriacetate (NTA)	0.27	0.40	46	0.35	28	0.38	39	[69] cited in [24]
Carbofuran	0.42	0.41	-1.4	0.49	16	0.52	25	[29]
MCPB		0.42		0.55		0.57		
Bifenazate		0.42		0.57		0.58		
Benalaxyl		0.43		0.57		0.58		
Acetamiprid		0.43		0.48		0.50		
Paraquat		0.43		0.57		0.59		
Chlorpropham		0.44		0.58		0.60		
Ibuprofen	0.30	0.46	52	0.54	79	0.56	85	[39]
Acetochlor		0.46		0.55		0.56		

Milbemectin		0.47		0.62		0.63		
Metalaxyl-M		0.48		0.58		0.60		
Alachlor		0.49		0.61		0.62		
Pendimethalin		0.51		0.58		0.60		
Glyphosate	0.18	0.51	183	0.58	224	0.60	234	[42]
Atrazine		0.52		0.51		0.53		
Daminozide		0.57		0.54		0.55		
Linear correlation to degree of reductance ( $R^2$ )		0.14		0.53		0.70		
Mean absolute error (MAE)			49%		82%		111%	

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Table 5. Comparison of bacterial growth yields of organic chemicals of environmental concern under anaerobic conditions in g cell carbon (g substrate carbon)<sup>-1</sup>. Predictions were made using MTB, TEEM2, and MTB. The cell formulation for MTB and TEEM2 was taken to be C<sub>5</sub>H<sub>7</sub>O<sub>2</sub>N with a degree of reductance of 4.0. For ETTYM it was taken to be C<sub>5</sub>H<sub>10</sub>O<sub>3</sub>N with a degree of reductance of 4.2. The predictions were made under standard biochemical conditions (pH = 7). The entries are sorted from low to high predicted yield of MTB under nitrate-reducing conditions. For TEEM2 a weak positive correlation and for ETTYM and strong positive correlation exists between the degree of reductance of the compound and predicted yield (*R*<sup>2</sup> is shown). Values highlighted in bold are negative and thus not meaningful.

Compound	MTB	TEEM2	ETTYM	MTB	TEEM2	ETTYM
Unit	g cell carbon (g substrate carbon) <sup>-1</sup>					
Electron acceptor	NO <sub>3</sub> <sup>-</sup>			SO <sub>4</sub> <sup>2-</sup>		
Chlorothalonil	0.00	0.34	0.35	0.00	0.11	0.12
Pyrene	0.25	0.42	0.50	0.02	0.02	0.05
2,4-D	0.28	0.46	0.47	0.07	0.14	0.15
Dicamba	0.28	0.45	0.47	0.07	0.14	0.15
DDT	0.29	0.51	0.53	0.07	0.15	0.17
Anthracene	0.29	0.52	0.54	0.05	0.07	0.11
Phenanthrene	0.29	0.52	0.54	0.05	0.07	0.11
Azoxystrobin	0.32	0.46	0.47	0.07	0.12	0.14
Famoxadone	0.33	0.50	0.52	0.08	0.15	0.16
Ethylenediaminetetraacetate (EDTA)	0.33	0.34	0.41	0.07	<b>-0.06</b>	0.11
Metamitron	0.33	0.42	0.43	0.07	0.11	0.12
Benzoate	0.33	0.47	0.52	0.05	<b>-0.02</b>	0.10
Propyzamide	0.34	0.51	0.52	0.08	0.13	0.15
Cypermethrin	0.35	0.51	0.53	0.08	0.14	0.15
MCPA	0.35	0.50	0.52	0.09	0.14	0.15
Benzene	0.36	0.45	0.55	0.04	<b>-0.07</b>	0.07

2,4-DB	0.36	0.51	0.52	0.09	0.16	0.17
Naphthalene	0.37	0.53	0.56	0.07	0.07	0.12
Desmedipham	0.37	0.54	0.55	0.12	0.21	0.23
Phenmedipham	0.37	0.53	0.55	0.12	0.21	0.23
Pymetrozine	0.38	0.44	0.46	0.12	0.17	0.19
Iprodione	0.38	0.47	0.49	0.12	0.18	0.20
Mecoprop (MCP)	0.38	0.52	0.54	0.10	0.15	0.16
Nitrilotriacetate (NTA)	0.38	0.34	0.37	0.10	0.04	0.11
Carbofuran	0.40	0.47	0.51	0.06	0.01	0.08
MCPB	0.40	0.53	0.55	0.10	0.14	0.16
Bifenazate	0.41	0.55	0.57	0.12	0.18	0.20
Benalaxyl	0.41	0.55	0.57	0.10	0.14	0.16
Acetamiprid	0.41	0.47	0.48	0.12	0.15	0.17
Paraquat	0.42	0.55	0.57	0.11	0.15	0.16
Chlorpropham	0.43	0.57	0.59	0.15	0.22	0.24
Ibuprofen	0.44	0.52	0.59	0.11	0.13	0.14
Acetochlor	0.45	0.52	0.54	0.07	0.08	0.09
Milbemectin	0.46	0.60	0.62	0.13	0.18	0.19
Metalaxyl-M	0.47	0.57	0.58	0.14	0.18	0.19
Alachlor	0.47	0.59	0.61	0.14	0.18	0.20
Pendimethalin	0.50	0.57	0.58	0.22	0.26	0.28
Glyphosate	0.50	0.57	0.59	0.23	0.29	0.32
Atrazine	0.51	0.50	0.52	0.21	0.22	0.24
Daminozide	0.56	0.53	0.54	0.24	0.22	0.23
Linear correlation to degree of reductance ( $R^2$ )	0.14	0.49	0.67	0.00	0.01	0.01

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Table 6. Bacterial growth yields on simple carbon substrates under aerobic conditions predicted using MTB, TEEM2 and ETTYM in g cell carbon (g substrate carbon)<sup>-1</sup>. Where experimental data have been available, the predictions are evaluated with respect to these. The error and mean absolute error are shown. The experimental bacterial growth yields ( $Y^{EXP}$ ) were taken from Xiao and VanBriesen [24]. The predictions were done under standard biochemical conditions (pH = 7). For ETTYM the cell formulation is CH<sub>10</sub>O<sub>3</sub>N with a degree of reductance of 4.2. For TEEM2 and MTB it is C<sub>5</sub>H<sub>7</sub>O<sub>2</sub>N with a degree of reductance of 4.0. The entries are sorted based on the highest absolute error calculated for MTB.

Compound	$Y^{EXP}$	MTB	E(%)	TEEM2	E(%)	ETTY M	E(%)
Unit	g cell carbon (g substrate carbon) <sup>-1</sup>						
Formate	0.16	0.40	155%	0.22	40%	0.23	45%
Oxalate	0.07	0.00	-100%	0.14	95%	0.11	59%
Glycine	0.25	0.38	55%	0.35	41%	0.36	44%
Phenylacetic acid	0.48	0.35	-27%	0.48	0%	0.52	11%
Fumaric acid	0.32	0.24	-25%	0.36	12%	0.37	15%
Formaldehyde	0.47	0.58	23%	0.51	9%	0.53	13%
Malonate	0.24	0.29	23%	0.31	30%	0.29	23%
Citrate	0.37	0.29	-22%	0.35	-7%	0.35	-7%
Glyoxylate	0.22	0.27	21%	0.27	23%	0.28	26%
α-D-Fructose	0.51	0.61	20%	0.49	-3%	0.51	0%
Gluconate	0.53	0.62	18%	0.57	7%	0.58	10%
Sorbitol	0.55	0.47	-16%	0.53	-4%	0.54	-2%
α-D-Glucose	0.53	0.61	15%	0.49	-6%	0.51	-4%
Mannitol	0.55	0.47	-15%	0.53	-4%	0.54	-1%
α-Lactose	0.51	0.44	-14%	0.50	-2%	0.51	1%
Acetate	0.42	0.47	12%	0.41	-3%	0.45	8%
Lactate	0.49	0.45	-9%	0.46	-7%	0.47	-5%
α-D-Galactose	0.56	0.61	9%	0.49	-12%	0.50	-9%
Phenol	0.36	0.33	-8%	0.44	23%	0.53	48%

Malate	0.34	0.32	-7%	0.35	2%	0.36	5%
Succinate	0.39	0.37	-4%	0.38	-3%	0.40	1%
Propionate	0.48	0.50	3%	0.47	-1%	0.53	10%
Pyruvate	0.38	0.39	3%	0.39	4%	0.40	7%
Xylose	0.49	0.48	-2%	0.58	17%	0.59	21%
Tartrate	0.28	0.27	-2%	0.35	26%	0.36	30%
Glycerol	0.62	0.62	0%	0.56	-9%	0.58	-6%
Mean absolute error (MAE)			23%		15%		16%

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Table 7. Sensitivity analysis of all three methods. The average sensitivity,  $\bar{S}_i$ , and standard deviation shown in brackets. The degree of reductance does not affect the predictions of MTB. The standard cell formula for ETTYM is  $C_5H_{10}O_3N$  ( $\gamma_s = 4.2$ ,  $\sigma_c = 0.45$  gC (g cell dw) $^{-1}$ ) and for MTB and TEEM2 it is  $C_5H_7O_2N$  ( $\gamma_s = 4$ ,  $\sigma_c = 0.53$  g cell carbon (g cell dw) $^{-1}$ ). The alternative cell formulae used were:  $C_5H_{8.33}O_{0.8}N$  ( $\gamma_s = 4.74$ ,  $\sigma_c = 0.63$  g cell carbon (g cell dw) $^{-1}$ ),  $C_{4.1}H_{6.8}O_{2.2}N$  ( $\gamma_s = 3.85$ ,  $\sigma_c = 0.47$  g cell carbon (g cell dw) $^{-1}$ ).

Parameter	Method	Relative change in parameter value	Average sensitivity, $\bar{S}_i$ (standard deviation)
$Y_{ATP}$	MTB	+20%	0.56 (0.073)
	TEEM2		0.51 (0.016)
	ETTYM		0.49 (0.006)
	MTB	-20%	0.66 (0.068)
	TEEM2		0.61 (0.016)
	ETTYM		0.59 (0.0046)
$\Delta G_r^{\circ'}$	MTB	+50%	0.010 (0.090)
	TEEM2		-0.004 (0.25)
	ETTYM		0.006 (0.21)
	MTB	-50%	0.015 (0.080)
	TEEM2		0.035 (0.17)
	ETTYM		0.024 (0.15)
C <sub>5</sub> H <sub>10</sub> O <sub>3</sub> N			
Carbon content	MTB	-14%	0.59 (0.12)
	TEEM2	-14%	0.71 (0.02)
Degree of reductance	TEEM2	5%	-2.05 (0.05)
C <sub>5</sub> H <sub>7</sub> O <sub>2</sub> N			
Carbon content	ETTYM	17%	0.61 (0.03)
Degree of	ETTYM	-5%	-2.14 (0.12)

reductance			
$C_5H_{8.33}O_{0.8}N$			
Carbon content	MTB	19%	0.53 (0.12)
	TEEM2	19%	0.81 (0.40)
	ETTYM	39%	0.30 (0.01)
Degree of reductance	TEEM2	19%	0.81 (0.40)
	ETTYM	13%	0.89 (0.04)
$C_{4.1}H_{6.8}O_{2.2}N$			
Carbon content	MTB	-12%	0.60 (0.12)
	TEEM2	-12%	-0.67 (0.57)
	ETTYM	3%	-1.77 (0.15)
Degree of reductance	TEEM2	-4%	-2.17 (1.86)
	ETTYM	-8%	0.62 (0.05)

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Table 8. Predicted yields, measured CO<sub>2</sub>, non-extractable residues (NER), amino-acids, and predicted formation of soil organic matter (SOM) (biogenic NER) from Eq. (4) ( $SOM_{biogenic\ NER}$ ) (in brackets: result of Eq. (3), short-term experiments only). The data on measured CO<sub>2</sub> and NER formation were taken from Barriuso *et al.* [36] except where indicated otherwise. The data points are the maximum values reported for the experiments with the longest duration, if more than one experiment was reported. The entries are sorted from high to low predicted biogenic NER formation.

Compound	$Y_{MTB}^{EST}$	Measured CO <sub>2</sub>	Measured NER	Measured carbon label in amino acids	Predicted $SOM_{biogenic\ NER}$ using Eq. (4) ( $X_{biomass\ NER}$ using Eq. (3))
Unit	mol C <sub>bacteria</sub> (mol C <sub>substrate</sub> ) <sup>-1</sup>	% of applied labelled compound			
Glyphosate	0.51	80.1	8.8		28
Daminozide	0.57	59	25		24
Glyphosate <sup>1</sup>	0.51	50	30.4	11 – 12	17 (52)
MCPB	0.43	58	30		15
MCPA	0.38	67	30		15
Ibuprofen <sup>2</sup>	0.46	45	29.6	28	13 (38)
Mecoprop (MCP)	0.41	52	47		13
Desmedipham	0.39	46.4	55		11
Milbemectin	0.49	35	40		11
Metalaxyl-M	0.5	33	73		11
Cypermethrin	0.36	48	26		11

Propyzamid	0.36	48	27		11
2,4-D <sup>3</sup>	0.3	58	36	23	10 (24)
Metamitron <sup>4</sup>	0.35	49	41	15	10 (25)
2,4-DB	0.38	42.1	33.2		10
Benalaxyl	0.43	25	18.8		7
Famoxadone	0.34	32.2	51.4		7
Bifenazate	0.43	23	67.3		6
2,4-D	0.3	36	27.9		6
Carbofuran <sup>5</sup>	0.41	23.8	63		6
Phenmedipham	0.4	16.5	64.1		4
Pymetrozine	0.4	15	61		4
Azoxystrobin	0.34	14	24		3
Acetamiprid	0.43	9.6	32.3		3
Iprodione	0.41	5	40		1
Pendimethalin	0.5	2.4	10		1
Chlorothalonil	0	15	54		0

862 <sup>1</sup>[42] <sup>2</sup>[39] <sup>3</sup>[38] <sup>4</sup>[43] <sup>5</sup>[64]

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Authors' response for

Microbial growth yield estimates from thermodynamics and its importance for degradation of pesticides and formation of biogenic non-extractable residues.

(Manuscript ID SQER-2017-0051)

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We greatly appreciate the feedback we received from the anonymous reviewers. In this document, we have addressed all the comments made (blue font colour) and we have listed the revisions made to the manuscript in quotation marks. The revisions made to sentences are marked in blue font colour. In the manuscript, the revisions are included in the same blue font colour.

Referee: 1

Comments to the Author

A thermodynamically based approach is presented for modelling the degradation of chemicals and the formation of non extractable residues. The manuscript extends an approach that is submitted to another journal and is still under review.

Although the authors complete their conclusions section on a positive note, it is difficult to see an added value of the approaches presented in this manuscript. Basically, the main conclusion is that other factors than thermodynamics are more important in modulating degradation and bound residue formation. It is especially surprising in this respect to see that at the end, setting a default value to the Gibbs free energy of formation is concluded to be a suited approach as the final results are independent of this parameter. In addition to this key issue, there is the problem of the underlying conceptual manuscript still being under review. Thus, there is currently no real basis for the key model used in this contribution.

We thank the reviewer for the comment. We do agree with the reviewer that the conceptual manuscript should be accepted (and published online) before the publication of the present manuscript. Adding to this, we think that the reviewer has misunderstood the conclusion regarding the Gibbs free energy. Our finding was not that other factors than thermodynamics are important in modulating degradation and bound residue formation (although they surely are). The finding is that the Gibbs free energy of formation of the *chemical of interest* (i.e. pesticide) can be set to 0 kJ mol<sup>-1</sup> if no reliable data is available without affecting the outcome of the MTB method more than a few percent. This is because the majority of energy in the reaction comes from the formation of CO<sub>2</sub> and H<sub>2</sub>O – especially if the pesticide contains many C and H atoms.

To make this clearer we changed the sentence in the conclusions accordingly. It now reads (lines 541-543):

“All methods showed low sensitivity to variations in the Gibbs energy of formation of the organic chemicals because most of the Gibbs energy of reaction stems from the formation of the oxidation products carbon dioxide and water.”

Referee: 2

Comments to the Author

Authors: Andreas Libonati Brock,\* Matthias Kästner, Stefan Trapp

Article: Microbial growth yield estimates from thermodynamics and its importance for degradation of pesticides and formation of biogenic non-extractable residues

Journal: SAR and QSAR in Environmental Research

Manuscript: SQER-2017-0051

This is a valuable contribution on the evaluation and/or prediction of non-extractable residues formation during (bio)degradation of pesticides and I would like to see it published. However, this manuscript has several major problems, some of which is of a formal nature, and needs major revision before it will be suitable for publication. The detailed description of each critical point is given in the following section, complemented with the suggestion for possible improvement.

This is an outstanding review, helpful, detailed, and careful. We are very grateful to this reviewer for the efforts made to improve our manuscript. Thank you! The contribution has been duly noted in the acknowledgements. The critical points given in the following section are all addressed.

Major problems:

1. This manuscript cannot and should not be published before reference 16, manuscript submitted by the same authors to Environmental Science and Technology (ES&T) journal, is accepted for publication and published on-line. Namely, manuscript submitted to ES&T and its content is of critical importance for the main objective of this study, comparison of MTB method with the existing yield estimation methods. Reference 16 is cited 15 times in the text of this manuscript which clearly demonstrates its critical relevance for this manuscript.

We do agree with the reviewer that the conceptual manuscript should be accepted (and published online) before the publication of the present manuscript. Reference 16 (in the revised manuscript it is reference 15) was submitted several months before this manuscript, but we had been given more time for revision. We will resubmit ref 16 earlier so that it is published soon (but of course we cannot guarantee this :-))

2. Materials and Methods section should be significantly reduced since to a large extent it duplicates materials and results that are already published in the open literature.

Lines 118-123 now reads:

"The bacterial growth yield prediction methods chosen for this study have a common basic approach: a stoichiometrically balanced redox reaction and the associated change of Gibbs free energy. This means that one can set up half-reactions describing the reduction of the targeted compound (be it xenobiotic or not), calculate the associated Gibbs free energy [34], and combine it with half-reactions of an appropriate electron acceptor."

Line 125 now reads:

"This approach has been shown for ETTYM and TEEM2 and here"

Line 128-131 has been changed and now reads:

"A detailed summary of the methods can be found in the Supplementary Information (SI) and in the original references. The element, energy and electron balances differ between the methods, thus a brief outline of the methods will be given in Table 1, in which the final equations used to calculate the growth yield are shown."

The section has been significantly reduced (for details on each sub-section in the section see below). Table S1 showing the final equations used for the growth yield predictions has been moved from the Supplementary Information to the manuscript. The table is now listed as Table 1. Consequently, the numbering of tables has been updated throughout the manuscript. Additionally, the amount of equations has been reduced and their numbering has also been updated throughout the text.

Specifically:

(i) sub-section "Microbial Turnover to Biomass (MTB)" covering 3.5 pages should be either deleted or reduced to a minimum (half page). All this is already described in reference 16 and only points relevant for the main objective of this study may be briefly presented in this sub-section.

The sub-section has been significantly reduced and now contains the following (lines 132-148):

"The Microbial Turnover to Biomass (MTB) method is presented in detail in Trapp *et al.* [15]. The method is based on the work of Diekert [18]. The maximum bacterial yield is determined from the nutritional value of the substrate (N) combined with the determination of bio-available electrons from the reaction. The nutritional value is the inverse of the yield and describes how much substrate is needed for the growth of bacteria [g substrate (g biomass)<sup>-1</sup>]. This is subdivided into a biomass yielding (anabolic) and energy yielding

(catabolic) part. The catabolic yield is determined from calculation of the Gibbs free energy released from the oxidation of the compound, the storage of this energy in ATP, and the bacterial growth yield on ATP.

Microbes cannot use all electrons to generate energy and thus the concept of bio-available electrons was introduced. [15]. Thus, energy and electron balances are implicitly considered. The anabolic yield is calculated from the carbon content in the compound (the carbon source) and in the bacterial cell [18], i.e. how many grams of cell can be produced from the carbon in the compound (only carbon availability is assumed to limit growth).

Further details and examples can be found in SI 1.1 and Trapp *et al.* [15]. “

The paragraph

“Gibbs energy of reaction is calculated using activities of the reactants and products assumed to be 1 M, except for H<sup>+</sup> which is assumed to be 10<sup>-7</sup> M (pH 7). The Gibbs free energy of reaction for non-standard conditions can be calculated as

$$\Delta G_r' = \Delta G_r^{o'} + R T \ln Q = \Delta G_r^{o'} + R T \ln \left( \frac{\prod_{i=1}^n [product]_i^p}{\prod_{i=1}^n [reactant]_i^r} \right) \quad (1)$$

where  $R$  is the ideal gas constant [8.314 J (K mol)<sup>-1</sup>],  $T$  is the absolute temperature [K],  $Q$  is the reaction quotient, and  $[product]$  and  $[reactant]$  are the activities of products and reactants, and  $p$  and  $r$  are their respective stoichiometric coefficients. From the equation itself it can be seen that when  $Q < 1$  the term is negative and when  $Q > 1$  the term is positive.”

has been moved to lines 182-191.

The bulk of the text has been moved to the supplementary information thus making it possible for the interested reader to easily locate a relevant summary of the method.

(ii) sub-section “Thermodynamic Electron Equivalent Model 2 (TEEM2)” covering 3 pages should be also either deleted or reduced to a minimum (half page). Content of this sub-section is covered in details in the original studies (references 18, 23, 24, 35) and there is no need for repetition here. Again, only a brief outline of points relevant for the main objective of this study may be given in this sub-section.

The sub-section has been significantly reduced and now contains the following (lines 151-162):

“In 1965 P. L. McCarty presented a thermodynamic model to estimate the maximal bacterial yield from a single substrate [17]. The method determines the yield on a given substrate from the Gibbs free energy released in the redox process. Since its inception it has been modified and expanded [34]. It was recently modified to better capture the observed lower yields associated with C1 compounds (i.e. methanol) and reactions involving oxygenases [22, 23]. It is based on electron and energy balances. The electron balance considers that the electrons provided by the substrate are used either in the synthesis of cell material (anabolism) or in energy generation (catabolism), and the energy balance states that the energy *captured* with a specific efficiency ( $\epsilon$ ) by the organism is used for bacterial growth. The energy capture efficiency,  $\epsilon$ , is a key parameter and is estimated from experimental data. Further details and examples are found in SI 1.2 and McCarty [23] and Rittmann and McCarty [34]. “



The bulk of the text has been moved to the supplementary information thus making it possible for the interested reader to easily locate a relevant summary of the method.

(iii) sub-section “Expanded Thermodynamic True Yield Prediction Model (ETTYM)” covering 2.5 pages is also a waste of valuable journal space and should be reduced to a minimum. Its content is already published in references 23 and 25. Outline briefly points relevant for the main objective of this study.

The sub-section has been significantly reduced and now contains the following (lines 165-173):

“The Expanded Thermodynamic True Yield Model is based on the work by McCarty and was presented in [22] and expanded in [24]. To increase the accuracy for the yield prediction on C1 compounds and substrates with low degrees of reduction, the authors proposed to include a carbon and a nitrogen balance and as a result thereof reformulate the electron and energy balance originally proposed by McCarty. The carbon balance describes that the carbon is either invested in cell synthesis or into other carbonaceous products. The nitrogen balance can be ignored if nitrogen is not limiting [24], hence, the yield can be calculated from an energy balance, carbon balance, and electron balance. For further details, the reader is kindly referred to SI1.3 and [22, 24].”

The bulk of the text has been moved to the supplementary information thus making it possible for the interested reader to easily locate a relevant summary of the method.

3. It seems that the authors are not familiar with the mathematical concept of logarithms. Logarithms are only defined for positive numbers, those larger than zero. Thus, there are no logarithms for negative numbers. Consequently, the statement “when  $Q < 0$  the term is negative” on page 8, line 172 is meaningless.

Thank you. This is an obvious oversight - what is meant is that when  $Q < 1 \Rightarrow \ln(Q) < 0$ . This has been corrected. Lines 190-191 now reads:

“when  $Q < 1$  the term is negative and when  $Q > 1$  the term is positive.”

4. page 8, lines 169 and 170: There is no such unit as degree Kelvin ( $^{\circ}\text{K}$ ). The units of absolute temperatures are Kelvin (K).

Thank you for the comment. This has been corrected on line 187.

5. Another major drawback of this manuscript are numerous linguistic problems. A lot of sentences in this manuscript have awkward structure and are in contrast to the elementary principles of English language. This made the reading of this manuscript a demanding and time-consuming task. This manuscript should have been checked and corrected by someone proficient in English before it was submitted for a review. In a separate section “Some Linguistic Problems or Errors” I have listed some of those linguistic problems and have suggested possible corrections. However, in addition to those suggested corrections, a thorough inspection of the revised manuscript by someone proficient in English is still needed.

We sincerely thank the reviewer for taking time to meticulously go through the manuscript. The manuscript has undergone a thorough inspection with the aid from a proficient native speaker from the USA. The changes made are listed in the following.

Lines 19-22: "Formation of microbial mass can be estimated from the microbial growth yield but experimental data is rare. Instead, we suggest using prediction methods for the theoretical yield based on thermodynamics."

Lines 23-24: "We have estimated the growth yield on 40 organic chemicals"

Lines 28-29: "Having the theoretical yield and using the released CO<sub>2</sub> as a measure for microbial activity,"

Lines 40-41: "The evaluation of biodegradation of organic chemicals of environmental concern is a big challenge for risk assessment and is subject to legislation and regulation."

Lines 43-45: "The assessment of biodegradability under environmental conditions is standardised by OECD testing guidelines, such as OECD Tests Nos. 306-309 used for"

Lines 47-48: "Transformation and biodegradation is mostly tested with <sup>14</sup>C or <sup>13</sup>C labelled parent compounds."

Lines 48-49: "Isotopes are particularly needed for assessment of non-extractable residues (NER; also called "bound residues") and tracing of unknown metabolites [8]."

Line 51: "Although there are several approaches for the reliable prospective assessment"

Lines 50-51: "Assessment of biodegradation is well established but may still have some pitfalls for various compounds [9-12]."

Lines 54-55: "The assessment of residue formation is still in its infancy and is not yet predictable."

Line 62-64: "Apparently, NER are mostly comprised of all types of residues and thus the assessment of the biogenic NER formation will also provide information about the amounts of the other types of NER formed."

Lines 65-66: "Compounds that are poor growth substrates and do not provide sufficient energy"

Lines 67-69: "Thus, the usability of the molecule, its energy content and suitability for anabolic processes has a profound impact on the evolutionary pressure to develop degradation pathways."

Lines 71-73: "In addition, thermodynamics can also be used to describe the potential growth of bacteria [17]."

Lines 73-74: "Essentially, bacterial growth is simplified and split into anabolic processes (energy demanding) and catabolic processes (energy producing) [18]."

Line 74-77: "The catabolic processes describe the energy released from the oxidation of a chemical or a substrate. In aerobic metabolism the oxidation product is usually CO<sub>2</sub> and H<sub>2</sub>O. The electrons derived from the oxidation are partly transferred (...)"

Lines 79-80: ", predominately from the formation of CO<sub>2</sub> and H<sub>2</sub>O [19]."

Lines 81-82: "The anabolic processes describe the substrate and energy use for the synthesis of new cell biomass."

Lines 89-91: "These estimates have previously been used for biotechnological purposes and for the estimation of activated sludge formation in waste water treatment processes,"

Lines 91-9: "The different growth yield estimation methods are based on a similar set of considerations [26, 27]."

Lines 106-108: "Therefore, yield estimation provides a tool for the assessment and prediction of biogenic NER formation in the degradation assessment of chemicals for regulatory purposes."

Lines 179-180: "Ammonia was taken as the sole nitrogen source so electrons for the assimilatory reduction of  $\text{NO}_3^-$  was not considered,"

Lines 193-194: "nitrate-reducing conditions ( $\text{NO}_3^-$  as the terminal electron acceptor) and sulphate-reducing conditions ( $\text{SO}_4^{2-}$  as the terminal electron acceptor)"

Line 202: ". However, for ETTYM this was not done"

Lines 210-211: "key input parameters were varied including  $Y_{ATP}$ ,"

Lines 226-227: "Bacterial growth yields have been experimentally assessed only for very few of the selected compounds."

Lines 229-231: "Moreover, ibuprofen and some polycyclic aromatic hydrocarbons (PAH) were also included as the bacterial growth yield has been experimentally determined for these."

Lines 232-233: "the methods were also evaluated using the growth yield determined for simple carbon substrates used in biotechnology."

Lines 233-234: "The compounds selected for comparison are based on the review and evaluation made by Xiao and VanBriesen [24]. "

Line 268: "In Table 4, the predicted bacterial growth yields under aerobic conditions are shown and"

Lines 274-276: ". The reason being the absence of carbon-hydrogen bonds. All other compounds except pyrene were predicted to have a bacterial growth yield of  $>0.3 \text{ g cell carbon (g substrate carbon)}^{-1}$  by all methods."

Lines 283-285: ". A weaker but still highly significant ( $p < 0.01$ ) correlation was found for TEEM2 ( $Y = 0.09 \gamma_s + 0.15$ ,  $R^2 = 0.53$ ), and a significant ( $p < 0.05$ ) but rather weak correlation was found for MTB ( $R^2 = 0.14$ )."

Lines 288-289: "and ETTYM overestimated by 56% and 25%, respectively. For glyphosate and anthracene,"

Line 299: "With sulphate as the terminal electron acceptor,"

Lines 300-301: "compared to the yields found under aerobic conditions (Table 5)"

Lines 301-302: "For both aerobic and nitrate-reducing conditions, the ranking of the estimated yields was close."

Lines 305-307: "MTB predicted 20 chemicals to have a yield of  $<0.1 \text{ g cell carbon (g substrate carbon)}^{-1}$ , while ten were predicted using TEEM2, and only four using ETTYM."

Lines 309-310: "The lack of experimental observations under a multitude of redox conditions makes assessment of the prediction accuracy unachievable."

Lines 318-319: "Subsequently, the work presented in [24] resulted in an even lower MAE when both the pH and related speciation, and oxygenase reactions were taken into account."

Lines (formerly) 320-322 were deleted: "(the ranking was identical from the lowest predicted yield and 12 places up; and the six substrates predicted to give the highest yield)"

Lines 323-325: "For MTB, these were formate, oxalate, and glycine; for TEEM2, these were oxalate, glycine, and formate; and for ETTYM, these were oxalate, phenol, and formate."

Lines 327-328: "with TEEM2 still giving the lowest MAE (9%)"

Line 342: "All methods are relatively sensitive to changes in  $Y_{ATP}$ , especially MTB."

Lines 344-446: "The chosen default value for xenobiotics of 5 g cell dw (mol ATP)<sup>-1</sup> ([18], for methanol) used in the MTB method does not lead to large errors (cf. Table 4)."

Lines 350-351: ". Conversely, a negative  $\Delta G_f^\circ$  which is made more negative leads to a decrease in the predicted yield (e.g., EDTA)."

Line 354: "Yield estimates for chemicals with few carbon atoms"

Line 355: "are especially sensitive to changes"

Lines 366-369: ". The effect of the cell formula in TEEM2 and ETTYM is not only on the energy costs related to synthesis [24], but also on the conversion to g cell carbon (g substrate carbon)<sup>-1</sup> in TEEM2. This is due to the degree of reduction of the cell ( $\gamma_s$ ) used in converting the units."

Line 376: "NER formation from the CO<sub>2</sub> produced during degradation experiments"

Lines 378-380: "Except for one compound (glyphosate, caused by the production of the metabolite aminomethylphosphonic acid (AMPA)), the predicted biogenic NER was smaller than the measured total NER."

Lines 383-384: "and pymetrozine the biogenic NER is suggested to make up less than 10% of the formed NER"

Line 402: "with some of them used earlier in [15]."

Lines 404-405: "The bacterial growth yield estimation methods are all developed to predict the true yield"

Lines 405-407: "The observed value is typically a net yield accounting only for the formation of new cell mass and removal of the parent compound"

Lines 407-409: "The difference between the two is that for the observed yield energy and carbon expenditure, due to non-growth purposes, are not captured"

Lines 409-410: "(e.g. energy spent on maintenance, formation of metabolites or soluble microbial products and extracellular polymeric substances)"

Line 411: “, unless a dynamic model was used for fit. Hence,”

Line 413-414: “If hardly degradable or insoluble metabolites are formed and rendered not bioavailable (as NER I or II),”

Lines 415-417: “can also result in a higher amount of energy being spent on maintenance leading to an observed yield lower than the predicted true yield. “

Lines 420-422: “, the intermediate AMPA accumulated, resulting in an observed yield much that was lower than the predicted yield (Table 4).”

Lines 424-425: “The presence of other sources of carbon or energy (mixed substrate use) also adds uncertainty to the observed value.”

Lines 431-432: “Under these conditions, microbes use most of the growth substrate just for maintenance [51].”

Lines 441-442: “This shows that the half-reaction approach using various electron acceptors used in ETTYM and TEEM2 can also be used with MTB”

Line 453: “the TEEM2 method was calibrated to the data in order to produce yield estimates”

Line 456: “proposed for the ETTYM method [24]. “

Line 458: “estimated it as zero due to the absence of C-H bonds (which points to the need for a”

Line 489: “. However, speciation of the substrate also has an effect on its Gibbs energy of formation.”

Line 490: “Similarly, the sensitivity of the energy capture efficiency parameter  $\epsilon$ ”

Line 496: “The peak in living biomass is usually after a few days to weeks [38, 39],”

Lines 508-510: “We suggested such a model in [8] and used it successfully, for the prediction of the NER formation from 2,4-D and ibuprofen with pre-estimated yield data [15].”

Line 511: “The data compiled by [36] give no hints into which form the NER are present”

Line 516: “done” deleted: “In degradation experiments with metatriton [43], glyphosate [42]”

Lines 518-519: “amino acids, fatty acids, metabolites, and parent compounds.”

Lines 525-526: “The method applied in this study provides process-based theoretical background that may be used to interpret NER data derived in degradation experiments.”

Lines 534-535: “while MTB performed better when estimating the yield on organic chemicals of environmental concern in general and in particular on pesticides.”

Line 808: “Predictions are evaluated using available experimental data.”

Line 826: “Values highlighted in bold are negative and thus not meaningful.”

6. lines 650-666 and 733-736: This extensive discussion and related conclusions are meaningless and should be deleted. Gibbs energy of formation can be easily calculated for any chemical by quantum-mechanical

methods. Thus, any limitation of group contribution methods is irrelevant and it also does not make sense to set the Gibbs energy of formation to zero.

Please allow that at this one issue we disagree with the reviewer. It may be possible to get exact values of Gibbs energy of formation, but personally we had some difficulties to find the values of several organic chemicals and we also found many conflicting data of delta G. We expect other users to have similar experiences. It is therefore a relevant message that small error or even missing delta G does not inhibit good estimates of the microbial yield of xenobiotics. Nonetheless, we shortened this section in 4.2 so that it is less extensive, and we deleted the lines in the conclusions, except one, which now reads (lines 541-543):

"All methods showed low sensitivity to variations in the Gibbs energy of formation of the organic chemicals because most of the Gibbs energy of reaction stems from the formation of the oxidation products carbon dioxide and water.

Section 4.2 *Sensitivity analysis* now reads (lines 462-478):

"All the methods were shown to be sensitive to the choice of cell formula but exhibited low sensitivity to variations of the formation energy,  $\Delta G_f^\circ$ , of the chemical of interest. All methods are based on the Gibbs energy of reaction and knowledge of the Gibbs energy of formation of the chemical of interest is needed. If the value has not been determined experimentally (e.g. [52]), it can be estimated using group contribution methods [53-58] (method [53] is implemented in the freely available ChemProp [59]), or by component contribution methods [60] (implemented in the free accessible database eQuilibrator [61]), or calculated using quantum mechanics [62]. For xenobiotics, the applicability of these estimation methods may be limited. Consequently, we also tested the sensitivity of Gibbs energy of formation of the xenobiotic substrate by setting this value to 0 kJ mol<sup>-1</sup> (Table S3). The MTB method has surprisingly low sensitivity. Compounds having a large negative Gibbs energy of formation (e.g. NTA, EDTA, and glyphosate) and few carbon-hydrogen bonds (6, 8, and 4, respectively) show a maximum deviation of around 20% from the predictions done with correct Gibbs energy of formation. Overall, the average deviation is only 6%. In comparison, TEEM2 and ETTYM have considerably higher average deviation (14% and 11%, respectively)."

#### Specific Problems or Errors:

1. line 54: References 13 and 15 are barely suitable to support the current status on modeling and assessment of biodegradation. Reference 13 is just a technical collection of published QSAR models without any critical evaluation of those models while reference 15 is only describing Biocatalysis/Biodegradation Database and the improved public access to this database. Thus, those references should be either replaced or, at least, amended by the recent critical review on the most relevant qualitative and quantitative models for estimating or evaluating biodegradation of organic chemicals and published in ACS Symposium Series:

A. Sabljic and Y. Nakagawa, Biodegradation and quantitative structure-activity relationship (QSAR), in Non-First Order Degradation and Time-Dependent Sorption of Organic Chemicals in Soil, W. Chen, A. Sabljic, S.A. Cryer and R.S. Kookana, eds, Book Series: ACS Symposium Series, American Chemical Society, Washington (DC), Volume 1174, 2014, pp. 57-84.

We thank the reviewer for the comment. After reading the suggested literature, we have included it as a reference and removed reference 15.

2. lines 115-117: Delete this sentence since it only repeats previous statement.

The sentence was deleted.

3. line 193: It seems that variable  $Y_{ATP}$  has not been defined.

Please note, that this sentence has been deleted from the manuscript and moved to the supplementary information. (Line SI 58)

Due to the changes made to the Material and Methods section,  $Y_{ATP}$  is now defined in Table 1.

4. line 266: "(McCarty 2007)", wrong format for this reference

The reference has been changed to the correct format:

"[23]"

Please note, that this sentence has been deleted from the manuscript and moved to the supplementary information.

5. line 352: From here the authors have started to number sections and sub-sections. Why? Previous sections and sub-sections are not numbered.

The un-numbered sections and subsections have now been numbered.

Line 39: "1. Introduction"

Line 117: "2. Materials and Methods"

Line 132: "2.1 Microbial Turnover to Biomass (MTB)"

Line 150: "2.2 Thermodynamic Electron Equivalent Model 2 (TEEM2)"

Line 164: "2.3 Expanded Thermodynamic True Yield Prediction Model (ETTYM)"

Line 177: "2.4 Conditions for comparison"

Line 222: "2.5 Chemicals of environmental concern; data"

Line 242: "2.6 Calculation of biogenic non-extractable residues"

6. lines 363-364: "SO<sub>4</sub><sup>2-</sup> is oxidised to H<sub>2</sub>S and HS<sup>-</sup>, NO<sub>3</sub><sup>-</sup> to N<sub>2</sub>, O<sub>2</sub> to H<sub>2</sub>O,"?! Do you mean reduced? Also "oxidized" is US spelling used in this manuscript.

Thank you for pointing this out. Yes, it is not oxidation but reduction as they are all electron acceptors.

"Oxidised" has been changed to "reduced" on line 198: "SO<sub>4</sub><sup>2-</sup> is reduced to H<sub>2</sub>S and HS<sup>-</sup>, NO<sub>3</sub><sup>-</sup> to N<sub>2</sub>, O<sub>2</sub> to H<sub>2</sub>O, C to CO<sub>2</sub>"

7. lines 409-442: The content of sub-section “2.3 Calculation of biogenic non-extractable residues” is described in details in reference 16 and only points relevant for the main objective of this study should be briefly presented in this sub-section.

Thank you for the comment. The sub-section has been revised and reduced one third. It now reads (Lines 243-264):

“Chemicals labelled with carbon isotopes ( $^{14}\text{C}$  or  $^{13}\text{C}$ ) allowed the flow of carbon to be tracked in the experimental system [38-41]. If the compound provides carbon to anabolism and cell synthesis, the labelled carbon will end up in microbial biomass and finally in the biogenic NER. Biogenic NER is not posing a risk to neither the environment nor human health [8]. When a substrate  $S$  is mineralized, the amount of biomass formed is yield times substrate,  $Y \times S$ , and the evolved  $\text{CO}_2$  is  $(1 - Y) \times S$  [15]. After the growth phase has stopped, the maximum ratio between biomass and  $\text{CO}_2$  and is thus

$$\frac{[X_{\text{biogenic NER}}]}{[\text{CO}_2]} = \frac{YS}{(1-Y)S} \text{ or } [X_{\text{biogenic NER}}] = \frac{Y}{1-Y} [\text{CO}_2] \quad (3)$$

where  $X_{\text{biogenic NER}}$  is the biomass making up the living biogenic NER. After the cessation of the growth phase, the microorganisms start to decay. The dead microorganisms are turned over in the microbial food chain and form new biomass,  $\text{CO}_2$  and soil organic matter (SOM) [30-33]. Then, the ratio between biogenic NER and  $^{13/14}\text{CO}_2$  becomes [15]

$$\frac{[SOM_{\text{biogenic NER}}]}{[\text{CO}_2]} = \frac{f \times Y}{(1-Y) + (1-f) \times Y} \text{ or } [SOM_{\text{biogenic NER}}] = \frac{f \times Y}{(1-Y) + (1-f) \times Y} [\text{CO}_2] \quad (4)$$

where  $SOM_{\text{biogenic NER}}$  is the non-living biogenic NER,  $f$  is the fraction of decaying biomass turned over into both living biomass and non-living SOM (0.5, [33]), and  $1-f$  is the fraction of label released as  $\text{CO}_2$ . Eq. (3) can be used to estimate NER formation during short-term experiments, whereas Eq. (4) holds for long-term experiments. It can be seen that a high mineralization and  $\text{CO}_2$  formation together with a high bacterial growth yield leads to a high formation of biogenic NER.”

8. lines 569-570: In Table 7 the predicted and measured biogenic NERs are not similar for 2,4-D. Clarify.

This requires some explanation. In the four cases where biogenic NER was reported (Karolina Nowak and her co-workers), the authors took the measured label in amino acids and multiplied with factor 2. This factor stems from the protein content (about 50%) in living microbes. However, as seen from the NER equations (now Eq. (3) and Eq. (4)), the biogenic NER is comprised of living and dead biomass. And for dead biomass, the factor 2 is not valid because sugars and fatty acids are metabolized much faster than amino acids. It can even be seen from the author's data that the application of the factor 2 to the amino acid fraction leads to false results because for ibuprofen and 2,4-D the fraction of biogenic NER is larger than the total NER. We therefore deleted the column "Reported biogenic NER" in Table 8, refer to the measured amino acids and changed the manuscript accordingly. As can be seen, the measured label in amino acid is always within the range of biogenic NER given by Eq. 3 (living biomass) and Eq. 4 (decayed biomass).

Now line 386-394:

“The experimental period for ibuprofen and 2,4-D (64 days) [38, 39], and glyphosate and metamitron (80 days) [42, 43] was shorter than the experiments reported in [36]. Eq. (3), which calculates living biomass  $X$  as biogenic NER, was additionally used to interpret these experiments. In these four studies, the carbon label found in amino acids was reported. For living microbes, about half of the carbon is in proteins. This fraction increases during decay and turnover of microbial biomass because proteins are the most stable fraction of the cells [30, 33, 15]. Except for glyphosate, the measured label in amino acids is within the range of biogenic NER predicted by Eqs. (3) and (4), and the measured total NER is greater.”



9. line 613: I would suggest to cite here also the recent review on bioavailability of xenobiotica in the soils environment.

A. Katayama, R. Bhula, G.R. Burns, E. Carazo, A. Felsot, D. Hamilton, C. Harris, Y.H. Kim, G. Kleter, W. Koerdel, J. Linders, J.G.M.W. Peijnenburg, A. Sabljic, R.G. Stephenson, D.K. Racke, B. Rubin, K. Tanaka, J. Unsworth and R.D. Wauchope, Bioavailability of xenobiotics in the soil environment, Reviews of Environmental Contamination and Toxicology 203 (2010), pp. 1-86.

Thank you for the suggestion. After reading the publication we agree with the reviewer. It is an excellent review and is now cited as ref 50 on line 430.

10. line 631: "any linearity"?! probably "there is no relationship"

Thank you for your comment. The sentence now reads (lines 448-450):  
"For all methods, the predicted growth yield is so small that there is no relationship between degree of reductance and bacterial growth yield."

11. line 649: References 55-59 should be deleted since those are not references for ChemProp estimation software, only reference 14 is relevant.

Thank you for the comment. References 55-59 are references for group contribution based methods. Lines 466-469 now read:  
"it can be estimated using group contribution methods [53-58] (method [53] is implemented in the freely available ChemProp [59]) or by component contribution methods [60] (implemented in the free accessible database eQuilibrator [61]) or calculated using quantum mechanics [62]."

12. line 650: Quantum mechanics is used to calculate and not to estimate the Gibbs energy of formation for chemicals.

Thank you for the comment. See above comment where 'calculated' has been added.

13. line 684: Reference 16 should be replaced by original reference(s) on "the peak in living biomass". Besides, reference 16 is not published yet.

Done. We cite here as example the experiments done with ibuprofen and 2,4-D (Refs 38, 39) (Line 496).

14. line 685: expression "is more likely applying to the data" is confusing and does not make sense

Thank you for the comment. The sentence has been revised and now reads (Lines 496-499):  
"and therefore we expect that Eq. (4) ( $SOM_{biogenic\ NER}$ ) is more appropriate for these experiments than Eq. (3) ( $X_{biogenic\ NER}$ ) as the majority of the living biomass has decayed and been incorporated into SOM after 100 days."

15. lines 685-687: What is implied by "this view"? No view is described in previous sentences of this subsection.

Lines 499-500 have been revised and now read:  
"Results obtained by Eq. (4) are smaller than the measured NER, which confirms the results of this equation."

See also comment 14.

16. lines 690-693: confusing long sentence

In regards to comment 37, the sentence has been changed to (lines 504-507):

"Such a correlation should not be expected since the processes leading to NER I, II and III are competing. If a pesticide is not degraded it can undergo aging and irreversible sorption (type I NER) and covalent binding of the parent compound or its metabolites (type II NER) [8]."

17. lines 700-701: What is the meaning of this sentence? How do you exploit the structure and energy of molecule by enzymatic pathways?

Thank you for the comment. The sentence has been revised and now reads (lines 514-515):

"Of course, the enzymatic pathways to facilitate the degradation and energy exploitation of the molecule also need to be present."

18. lines 716-718: This statement is not correct since the MTB method is already presented in reference 16.

Thank you for the comment. Changed to "applied". It now reads (lines 525-526):

"The method applied in this study provides process-based theoretical background that may be used to interpret NER data derived in degradation experiments."

19. lines 738-740: This conclusion is not correct since the MTB method is already presented in reference 16.

Thank your pointing this out. Changed to "applied". It now reads (lines 545-546):

"The approach applied in this study provides a theoretical foundation that can be used to predict biogenic NER formation during pesticide degradation experiments."

Some Linguistic Problems or Errors:

1. line 28: add comma, i.e. " for microbial activity, "

Thank you for pointing it out. The comma is added on line 29.

2. lines 56-57: correct sentence as "Recently, a novel approach was suggested for modelling the formation of biogenic residue [16] which can also shed light into the black box of NER."

Thank you for your suggestion. The sentence has been revised and now reads (lines 56-57):

"Recently, a novel approach was suggested for modelling the formation of biogenic residues [15] which can also elucidate the black box of NER."

3. lines 59-60: "but may also be formed by covalent bonding of metabolites (type II NER)" - Covalent bonding of metabolites to what? Why only metabolites, why is covalent bonding not possible for parent compound?

Yes, covalent bonding is possible for both parent compound and metabolites. The sentence was rewritten and now reads (lines 57-60):

"NER may be formed by sequestration or entrapment of parent compounds or metabolites in soils and sediments (*type I NER*), and also by covalent bonding to soil organic matter (*type II NER*)"

4. line 63: probably “comprise of all types”

Thank you for your suggestion. The sentence has been revised (line 62): “Apparently, NER are mostly comprised of all types of residues and”

5. lines 94-97: correct sentence as “Thermodynamic Electron Equivalent Model (TEEM2) developed by Perry L. McCarty [18, 24] and Expanded Thermodynamic True Yield Prediction Model (ETTYM) by Xiao and VanBriesen [23, 25] or their variations have been established and applied [29,30] for the estimation of bacterial growth yield on xenobiotics.”

Thank you for your suggestion. The sentence has been revised and the paragraph now reads (lines 93-99): “The Thermodynamic Electron Equivalent Model (TEEM2) developed by Perry L. McCarty [17, 23] and Expanded Thermodynamic True Yield Prediction Model (ETTYM) of Xiao and VanBriesen [22, 24] or their variations have been applied [28, 29] for the estimation of bacterial growth yield on xenobiotics. Both models have evolved towards an increased need of knowledge regarding the transformation pathways, metabolic processes and the electron and energy losses associated hereto in order to model specific growth of various organisms.”

6. line 98: expression “These models have both been moving” is meaningless in this sentence, maybe “Both models have been moving”

Thank you for your suggestion. The sentence has been revised (lines 96-97):

“Both models have evolved towards an increased need of knowledge regarding the transformation pathways,”

7. lines 104-105: correct text as “potential without the need for information on the pathways as this is rarely known for the majority of chemicals of environmental concern.”

Thank you for your suggestion. The sentence has been revised as suggested (lines 102-104):

“In MTB we proposed a simple method to predict just the minimum bacterial growth yield potential without the need for information on the pathways, as this is rarely known for the majority of chemicals of environmental concern.”

8. line 106: correct as “the microbial growth and decline are coupled”

Thank you for your suggestion. The sentence has been revised (line 105):

“Furthermore, microbial growth and decline are coupled to the formation of soil organic”

9. lines 110-111: correct text as “to thoroughly compare the recently introduced MTB method with other yield estimation methods”

Thank you for your suggestion. The sentence has been revised as suggested (lines 109-110):

“The objectives of the present study are i) to thoroughly compare the recently introduced MTB method with other yield estimation methods;”

10. line 128: “straightforward”? probably “direct”

The word has been deleted and the sentence now reads (lines 122-124):

“and combine it with half-reactions of an appropriate electron acceptor (e.g.  $O_2$ ,  $NO_3^-$ ,  $SO_4^{2-}$ ,  $Fe^{3+}$ ,  $Mn^{4+}$ , or even  $CO_2$  etc.) allowing for assessment of the bacterial growth yield under a multitude of redox conditions.”

11. line 129: Set as new sentence, i.e. “of redox conditions. Here we only look at  $O_2$ ,  $NO_3^-$  and  $SO_4^{2-}$ .”

Thank you for your suggestion. The sentence has been revised as suggested (lines 124-125):

“of redox conditions. Here we only look at  $O_2$ ,  $NO_3^-$ , and  $SO_4^{2-}$ .”

12. lines 160-161: correct as “the Gibbs free energy of reaction”

Thank you for your suggestion. The sentence has been revised as suggested. Note that the sentence has been moved from the manuscript to the supplementary information (Line SI 59).

13. lines 237 and 240: There is a mismatch between “(catabolism).” and “where”. The sentence is ending (i.e. period) on line 237 and new sentence is starting with the lower case letter on line 240. Analogous problem on lines 251 and 254.

The periods are removed. Note that the sentence has been moved from the manuscript to the supplementary information.

14. lines 246 or 247: Period is missing at the end of this sentence.

Thank you for the comment. A period has been added. Note that the sentence has been moved from the manuscript to the supplementary information.

15. line 250: “that all the energy”?! probably “that the whole energy”

Thank you for the comment. I have changed the sentence to read (correction in bold):

“The energy balance states that the energy *captured* by the organism from the redox reaction is used for bacterial growth.”

Note that the sentence has been moved from the manuscript to the supplementary information (Line SI 133).

16. lines 310-312: This sentence starts with equation which does not make sense. Also “Where” should be “where” since this is not the beginning of sentence.

The period in line has been removed and “Where” is correctly changed to “where”.

Note that the sentence has been moved from the manuscript to the supplementary information.

17. line 322: “Where” should be “where” since this is not the beginning of sentence.

Thank you for your suggestion. The sentence has been revised as suggested.

Note that the sentence has been moved from the manuscript to the supplementary information.

18. lines 342-344: This sentence also starts with equation which does not make sense or period on line 341 must be deleted.

Thank you for pointing this out. The period has been deleted.

Note that the sentence has been moved from the manuscript to the supplementary information.

19. lines 347-349: This sentence does not make sense at all. If electrons are only diverted to the reduction of the terminal electron acceptor (e.g.  $O_2$ ) and the synthesis of new cell material, and the electron donor is also the carbon source, the equation reduces to (where  $\gamma_{CO_2} = 0$ )

Thank you for bringing this to our attention. The sentence has been changed to:

“If the electron donor is also the carbon source and the electrons are used only for reduction of the terminal electron acceptor and synthesis of new cell material, the equation reduces to (remember,  $\gamma_{CO_2} = 0$ )”

I hope it is easier to make sense of it now. Note that the sentence has been moved from the manuscript to the supplementary information (Lines SI 226-228).

20. line 363: “oxidized” is US spelling used in this manuscript  
This has been corrected to “reduced” (line 198) – also see comment 6 under Specific Problems or Errors:

21. line 544: correct “MTB at least.”  
Since MTB is the method least sensitive to changes in the cell formulation, the sentence has been changed to (lines 365-366):  
“While all methods are sensitive to the cell formula used (Table 7), MTB is the method least affected.”

22. line 587: delete “experimentally” since it is redundant  
Thank you for your suggestion. Yes, it is clearly a pleonasm and it has been deleted (line 405):  
“The observed value is typically a net yield accounting only for the formation of new cell”

23. line 594: correct as “prediction methods assume a complete degradation of compound.”  
Thank you for your suggestion. The sentence has been revised on line 412:  
“Additionally, the prediction methods assume a complete degradation of the compound.”

24. lines 610-611: correct as “other carbon sources, and here the MTB predicted yields are very close to the experimentally determined values.”  
Thank you for your suggestion. The sentence has been revised as suggested (lines 426-428):  
“great care was taken in the experimental setup to minimise confounding factors due to other carbon sources, and here the MTB predicted yields are very close to the experimentally determined values.”

25. line 612: correct text as “The observed differences might also be attributed to their high hydrophobicity and”  
Thank you for your suggestion. The sentence has been revised as suggested (lines 429-430):  
“The observed differences might also be attributed to their high hydrophobicity and limited bioavailability”

26. line 616: correct text as “reactions for PAHs in TEEM2, its errors were higher than for MTB.”  
Thank you for your suggestion. The sentence has been revised as suggested (lines 433-434):  
“(putative) oxygenase reactions for PAHs in TEEM2, its errors were higher than for MTB.”

27. lines 618-619: correct text as “yield on carbofuran taking oxygenase reactions into account as suggested in [29].”  
Thank you for your suggestion. The sentence has been revised ( lines 435-436):  
“measured bacterial yield on carbofuran by taking oxygenase reactions into account as suggested in [28].”

28. lines 620-621: correct text as “Under sulphate-reducing conditions the predicted bacterial yields were much lower than the predicted yields under aerobic conditions, which can be expected”  
Thank you for your suggestion. The sentence has been revised as suggested (lines 437-439):  
“Under sulphate-reducing conditions the predicted bacterial yields were much lower than the predicted yields under aerobic conditions, which can be expected considering the lower”

29. line 623: correct “An interesting observation”  
Thank you for your suggestion. The sentence has been revised as suggested (line 440):

“An interesting observation was that the decrease in yield”

30. lines 626-627: correct text as “the majority of redox reactions might not be sufficient to fuel bacterial growth.”

Thank you for your suggestion. The sentence has been revised as suggested (lines 444-445):

“the energy released from the majority of redox reactions might not be sufficient to fuel bacterial growth”

31. line 639: “Where” should be “While”

Thank you for your suggestion. The sentence has been revised as suggested (line 457):

“While ETTYM and TEEM2 both overestimated the yield for oxalate,”

32. line 646: correct text as “reaction, knowledge of the Gibbs energy of formation of chemical is needed.”

Thank you for your suggestion. The sentence has been revised see comment 6 under Major Problems (lines 464-464):

“All methods are based on the Gibbs energy of reaction and knowledge of the Gibbs energy of formation of the chemical of interest is needed”

33. lines 670-671: correct text as “either due to the low energy associated with the oxidation of substrate (e.g. formate or formaldehyde) or the low energy associated with the reduction of electron acceptor”

Thank you for your suggestion. The sentence has been revised (lines 481-484):

“either due to the low energy associated with the oxidation of substrate (e.g. formate or formaldehyde), or the low energy associated with the reduction of the electron acceptor (e.g.  $\text{SO}_4^{2-}$ ). “

34. lines 685-687: modify this sentence as follows “Results obtained by Eq. (22) are smaller than the measured NER, which confirms this view.”

Thank you for your suggestion. The sentence has been revised (lines 499-500). See also comment 15 under Specific Problems or Errors:

“Results obtained by Eq. (4) are smaller than the measured NER, which confirms the results of this equation.”

35. lines 687-688: modify this sentence as follows “For daminozide, chemical with the highest predicted yield, calculated biogenic NER and measured NER are almost equal.”

Thank you for your suggestion. The sentence has been revised (lines 500-502):

“For daminozide, the chemical with the highest predicted yield, the calculated biogenic NER and measured NER are almost equal. “

36. line 689: modify text as “The inspection of Table 7 gives no significant correlation”

Thank you for your suggestion. The sentence has been revised and now reads (lines 503-504):

“The examination of Table 8 gives no significant correlation between measured total NER and predicted  $X_{\text{biogenic NER}}$  or  $SOM_{\text{biogenic NER}}$ .”

37. lines 690-693: modify this sentence as follows “Such a correlation should not be expected since the processes leading to NER I, II and III are competing and if pesticide is not degraded it can undergo aging, irreversible sorption and covalent binding of parent compound or its metabolites [8].”

Thank you for your suggestion. The sentence has been revised and now reads (lines 504-507):

“Such a correlation should not be expected since the processes leading to NER I, II and III are competing. If a pesticide is not degraded it can undergo aging and irreversible sorption (type I NER) and covalent binding of the parent compound or its metabolites (type II NER) [8].”

38. line 702: add comma after reference 46

Thank you for pointing this out, see comment 39 below

39. line 704: correct text "stable carbon or nitrogen isotope ( $^{13}\text{C}$  or  $^{15}\text{N}$ )"

Thank you for your suggestions. The sentence has been revised and now reads (lines 517-518)

"the formation of biogenic NER was investigated by tracking the distribution of stable carbon or nitrogen isotope ( $^{13}\text{C}$  or  $^{15}\text{N}$ ) in  $\text{CO}_2$ "

40. line 707: probably "The experimentally determined values"

Thank you for the comment. The sentence now reads (lines 519-521):

"Experiments of this kind are very helpful to discriminate between the various types of NER and to validate our biogenic NER estimation approach."

41. line 709: "although" does not make sense here since both stated results are positive

The sentence has been deleted as the paragraph has been revised.

42. lines 710-714: Another example of long and confusing sentence. "together with the release of  $\text{CO}_2$ " does not make any sense here while expression "to in fact be" is meaningless and cannot be combined with "surmised".

Thank you for the comment. Yes, we agree, the sentence is too long and is barely readable in one breath.

The sentence has been tidied up and is now (lines 521-524):

"Shrestha *et al.* [11] observed that the formation of NER occurred simultaneously with the degradation and release of  $\text{CO}_2$ . This shows the coupling of the formation of NER to microbial activity, and to the growth and decay of biomass."

43. lines 718-719: "dedicated experiments"?! maybe "targeted experiments"

Yes, targeted experiments done by dedicated researchers. This has been corrected on line 527-528:

"Before routine application though, further confirmation by targeted experiments is still needed"

44. lines 722-723: correct this sentence as "The MTB method was compared with two widely used bacterial growth yield estimation methods, TEEM2 and ETTYM."

Thank you for your suggestion. The sentence has been revised as suggested (lines 531-532):

"The MTB method was compared with two widely used bacterial growth yield estimation methods, TEEM2 and ETTYM."

45. line 727: correct text as "to electron acceptors other than oxygen,"

Thank you for your suggestion. The sentence has been revised (lines 536-537): "the MTB approach can be expanded to electron acceptors other than oxygen, like sulphate and nitrate."

46. line 730: delete "which is" since it is redundant

Thank you for your suggestion. The sentence has been revised (line 539):

"changes in  $Y_{ATP}$ , an uncertain parameter."

47. line 730: correct as "are also sensitive to"

Thank you for your suggestion. The sentence has been revised (lines 539-540):

"TEEM2 and ETTYM are also sensitive to changes in the cell formula"

48. line 731-732: correct the first part of this sentence as "All methods were insensitive with respect to the imprecise data on Gibbs energy of formation"

Thank you for your suggestion. The sentence has been revised (lines 540-543):

“All methods showed low sensitivity to variations in the Gibbs energy of formation of the organic chemicals because most of the Gibbs energy of reaction stems from the formation of the oxidation products carbon dioxide and water.

49. line 748: probably “spreadsheet”

Thank you for pointing this out. This has been corrected (line 556). Apparently my MS Word automatically changes it to “spread sheet.”

50. line 749: correct text as “on request from the first author.”

This has been corrected. Line 557 now reads:

“on request from the first author.”

### Other changes

Other changes made to the manuscript are listed here.

Line 773: “and Figures” deleted.

Lines 817-818 (deleted “anoxic”): “Comparison of bacterial growth yields of organic chemicals of environmental concern under anaerobic conditions in g cell carbon (g substrate carbon)<sup>-1</sup>.”

Lines 822-823: “The entries are sorted from low to high predicted yield of MTB under nitrate-reducing conditions.”

In sub-section 3.3 *Prediction of biogenic NER formation based on the predicted growth yields* the following sentence was deleted as it was an unnecessary repetition:

“High CO<sub>2</sub> formation as a measure for microbial activity typically resulted in the prediction of a large biogenic NER pool.”

To reflect the contributions made by the anonymous reviewers, the acknowledgements now read (lines 550-555): “This research project was financially supported by the Technical University of Denmark and the Helmholtz Centre for Environmental Research UFZ. We thank Fabio Polesel, Carson Odell Lee, and Ulrich Bay Gosewinkel for valuable suggestions and discussions. We also wish to acknowledge the comments and suggestions provided by the anonymous reviewers which helped to improve the manuscript.”

After having re-examined the tables, it appears that the values in column 2 ( $Y^{EXP}$ ) in Table 6 are incorrectly listed. They have somehow shifted during the transfer from the spreadsheet to the text document. The values in the other columns are correct, hence, the analysis and conclusions made based on Table 6 are all correct and unchanged.

Moreover, wrong values for  $Y_{ATP}$  (= 5) and C-H-bonds (= 6) for gluconate have been used. Instead the values should have been  $Y_{ATP}$  = 10 and C-H-bonds = 12. This has been changed in the supplementary information Table S2 and thus corrected in Table 6 and Table 7.

Additionally, the calculations made for the preparation of Table 7 also contain a minor calculation error. Instead of using  $Y_{ATP}$  = 10.5 for the calculation of the growth yield for TEEM2 and ETTYM methods, a value of  $Y_{ATP}$  = 10.5 × 0.8 = 8.4 has been used. Table 7 has been updated with correctly calculated values. The conclusions drawn on the original erroneous calculations are still valid.



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Consequently, Figure S1 has also been changed to reflect this. The changes do not change any conclusion made.

For Peer Review Only

## Supplementary Information

### Microbial growth yield estimates from thermodynamics and its importance for degradation of pesticides and formation of biogenic non-extractable residues

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The supporting information contains information regarding:

- i) S1. Summary of the methods used to estimate the microbial growth yields.
- ii) Table S1. Balanced half-reactions of the simple carbon substrates and the Gibbs free energy of reaction
- iii) Table S2. Gibbs free energy of formation, C-H bonds,  $Y_{ATP}$ ,
- iv) Table S3. The relative deviation from the estimated bacterial growth yield estimated under aerobic conditions when assuming a Gibbs free energy of formation of 0 kJ mol<sup>-1</sup>.
- v) Figure S1. The effect of cell formula on the yield estimate
- vi) References

18 pages; 4 tables; 1 figure

**S1. Summary of the methods used to estimate the microbial growth yields.**

A summary of the three methods to estimate the microbial growth yields is provided in the following.

***S1.1 Microbial Turnover to Biomass (MTB)***

The Microbial Turnover to Biomass (MTB) method presented in Trapp *et al.* [1] was based on the work by Diekert [2] where the maximum bacterial yield is determined from the nutritional value of the substrate ( $N$ ) combined with the determination of bio-available electrons from the reaction. The nutritional value describes how much substrate is needed per mass of bacteria [g substrate (g biomass)<sup>-1</sup>]. This is subdivided into a biomass yielding (anabolic) and energy yielding (catabolic) part. The nutritional value is the inverse of the yield, when the yield is defined as gram of cells per gram of substrate

$$N = N_{ana} + N_{cata} = \frac{1}{Y} = \frac{1}{Y_{ana}} + \frac{1}{Y_{cata}} \quad (S1)$$

$$\text{or } Y = \frac{Y_{ana} \times Y_{cata}}{Y_{ana} + Y_{cata}} \quad (S2)$$

where  $Y_{ana}$  is the anabolic yield,  $Y_{cata}$  is the catabolic yield, and  $Y$  is the bacterial growth yield [all in g bacteria dry weight (dw) (g substrate)<sup>-1</sup>].

***Catabolism***

The catabolic yield is determined from calculation of the Gibbs free energy released from the complete oxidation of the compound, the storage of this energy in ATP, and the bacterial growth yield on ATP

$$Y_{cata} = \frac{\frac{\Delta G^{0'}_{reaction}}{\Delta G^{0'}_{ATP} / \eta} Y_{ATP}}{M_s} \quad [\text{g cell dw (g substrate)}^{-1}] \quad (S3)$$

54

55  $\Delta G^{0'}_{ATP}$  is the Gibbs free energy of hydrolysis of adenosine triphosphate (ATP) [30.53  
 56 kJ mol<sup>-1</sup> [3]]. Bacteria have approximately 40% efficiency in the ATP energy gain ( $\eta$ )  
 57 [2, 4]. Taking this into account, the energy needed to synthesise one mole of ATP is  
 58 approximately -80 kJ (mol ATP)<sup>-1</sup>.  $Y_{ATP}$  is the bacterial yield on ATP [g cell dw (mol  
 59 ATP)<sup>-1</sup>] and its estimation is explained later.  $\Delta G^{0'}_{reaction}$  is the Gibbs free energy of  
 60 reaction,  $\Delta G^{0'}_{reaction} = \Delta G^{0'}_{products} - \Delta G^{0'}_{reactants}$ , at biological standard state conditions  
 61 [kJ mol<sup>-1</sup>].

62 The energy released from the reaction is assumed to not be fully available for  
 63 the microorganism and thus the concept of bio-available electrons was introduced [1].

64 The Gibbs free energy of reaction is related to the redox potential ( $\Delta E$ ) of the  
 65 reaction through the Nernst equation

$$66 \quad \Delta G = -n \times F \times \Delta E \quad (S4)$$

67

68 where  $n$  is the number of electrons transferred (sum of the change of oxidation status of  
 69 the carbon atoms in the substrate during oxidation), and  $F$  is the Faraday constant  
 70 [=96,485 C mol<sup>-1</sup>].

71 When C-H compounds are oxidised to CO<sub>2</sub> and H<sub>2</sub>O, the transferred electrons  
 72 are readily available for energy gain (2e<sup>-</sup> per C-H bond). Thus, the number of bio-  
 73 available electrons is  $n_{bio} \geq \text{no. of C-H bonds} \times 2$ . Therefore; the minimum bio-available  
 74 energy can be calculated from the number of C-H bonds.

$$75 \quad \Delta G^{0'}_{bioavailable} = \frac{n_{bio}}{n} \Delta G^{0'}_{reaction} \quad (S5)$$

76

77 The “bio-available energy corrected”  $Y^*_{cata}$  can then be formulated as

$$78 \quad Y^*_{cata} = \frac{\frac{\Delta G^{0'}_{bioavailable}}{(\Delta G^{0'}_{ATP} / \eta)} Y_{ATP}}{M_S} \quad (S6)$$

S3

79

80 Finally, as  $Y_{ATP}$  has been shown to be variable and dependent upon the substrate (and  
81 substrate concentration), e.g. [5, 6], a simple set of rules have been proposed in order to  
82 define this parameter. The rules are based on the  $\langle \text{CH}_2\text{O} \rangle$ -“sugar-structural-similarity”  
83 of the compound [1]:

84

- 85 • If  $\langle \text{CH}_2\text{O} \rangle$  -  $\langle \text{C}_2\text{H}_4\text{O}_2 \rangle$  is present in the molecule, or if oxygen is missing,  $Y_{ATP} = 5$   
86 g cell dw (mol ATP)<sup>-1</sup>
- 87 • If  $\langle \text{C}_3\text{H}_6\text{O}_3 \rangle$  -  $\langle \text{C}_4\text{H}_8\text{O}_4 \rangle$  is present in the molecule,  $Y_{ATP} = 7.5$  g cell dw (mol  
88 ATP)<sup>-1</sup>
- 89 • If  $\langle \text{C}_5\text{H}_{10}\text{O}_5 \rangle$  -  $\langle \text{C}_6\text{H}_{12}\text{O}_6 \rangle$  is present,  $Y_{ATP} = 10$  g cell dw (mol ATP)<sup>-1</sup>

90

91 Moreover, compounds not containing any “sugar-like” structure are assumed to have a  
92  $Y_{ATP} = 5$  g cell dw (mol ATP)<sup>-1</sup>. The oxygen atoms in N-O, S-O, P-O bonds and  
93 carboxylic groups are not counted [1].

94

#### 95 *Anabolism*

96 The anabolic yield,  $Y_{ana}$ , is calculated from the carbon content in the compound (the  
97 carbon source) and in the bacterial cell [2], i.e. how many grams of cell can be produced  
98 from the carbon in the compound (only carbon availability is assumed to limit growth)

$$99 \quad Y_{ana} = \frac{n_C M_C}{\sigma_C M_S} \quad [\text{g cell dw (g substrate)}^{-1}] \quad (S7)$$

100

101 where  $\sigma_c$  is the fraction of carbon in dry cell (here taken as  $\sim 0.53$  g C (g cell dw)<sup>-1</sup>, with  
102 the cell formula  $\text{C}_5\text{H}_7\text{O}_2\text{N}$  [7] but suggested to be  $0.5$  g C (g cell dw)<sup>-1</sup> in [2]),  $M_C$  is the

molar mass of carbon ( $12.01 \text{ g mol}^{-1}$ ),  $M_S$  is the molar mass of the substrate [ $\text{g mol}^{-1}$ ], and  $n_c$  is the number of carbon atoms in the substrate [ $\text{mol C (mol substrate)}^{-1}$ ].

To convert the yield from  $\text{g cell dw (g substrate)}^{-1}$  to  $\text{g cell carbon (g substrate carbon)}^{-1}$  the conversion can be determined from

$$f_{g/c} = \frac{1}{Y_{ana}} = \frac{\sigma_C}{M_C} \times \frac{M_S}{n_C} \quad (\text{S8})$$

where  $f_{g/c}$  has the unit  $(\text{g cell carbon (g substrate carbon)}^{-1}) (\text{g cell dw (g substrate)}^{-1})^{-1}$ .

### *S1.2 Thermodynamic Electron Equivalent Model 2 (TEEM2)*

In 1965 P. L. McCarty presented a thermodynamic model to estimate the maximal bacterial yield [8]. Since its inception it has been modified and expanded [9]. As the original model was found unable to capture the observed lower yields associated with C1 compounds (i.e. methanol) and for reactions involving oxygenases [4], it was recently modified [7]. The yield is calculated from an energy balance and an electron balance.

#### *Electron balance*

The electron balance states that the electrons provided by the substrate are used either in synthesis of cell material (anabolism) or in energy generation (catabolism)

$$f_s^0 + f_e^0 = 1 \quad (\text{S10})$$

where  $f_s^0$  is the fraction of electrons diverted for synthesis and  $f_e^0$  is the fraction of electrons diverted for energy generation (used to reduce the electron acceptor). The bacterial yield is equal to the fraction of electrons that are diverted to cell synthesis and formation of new biomass ( $f_s^0$ ). The yield can be calculated as  $\text{g cell carbon per g substrate carbon}$  (which is the very same as  $\text{mol C per mol C}$ ) from the degree of

128 reductance of the cell carbon ( $\gamma_c$ ) and the degree of reductance of the substrate carbon  
129 ( $\gamma_s$ ) [10].

$$130 \quad Y_{C/C} = f_s^0 \frac{\gamma_s}{\gamma_c} \quad (S11)$$

131

### 132 *Energy balance*

133 The energy balance states that the energy captured by the organism from the redox  
134 reaction is used for bacterial growth

$$135 \quad -f_e^0 \epsilon \Delta G_r^{0'} = f_s^0 \Delta G_s^{0'} \quad (S12)$$

136

137 where  $\Delta G_s^{0'}$  is the Gibbs free energy for synthesis [ $\text{kJ mol}^{-1}$ ],  $\Delta G_r^{0'}$  is the Gibbs free  
138 energy released from the redox reaction [ $\text{kJ mol}^{-1}$ ], and  $\epsilon$  is the energy capture  
139 efficiency.  $\epsilon$  is a key parameter and is estimated from experimental data. In [7] a best fit  
140 between predicted and experimental values was observed when it was set to 0.37 – i.e.  
141 37% of the energy released from the redox reaction is captured by the bacterium to be  
142 used in synthesis. It has generally been found to vary as a function of the growth (is it  
143 autotrophic or heterotrophic) [4, 9].

144 The Gibbs free energy of reaction is defined as

$$145 \quad \Delta G_r^{0'} = \Delta G_a^{0'} - \Delta G_d^{0'} - \frac{q}{p} \Delta G_{xy}^{0'} \quad (S13)$$

146

147 where the subscript  $a$  denotes the electron acceptor while  $d$  denotes the electron donor  
148 of the redox reaction,  $\Delta G_{xy}^{0'}$  is the reduction potential of NADH/NAD<sup>+</sup> oxidation (equal  
149 to  $-219.2 \text{ kJ mol}^{-1}$  [7],  $q$  is the number of oxygenase reactions [oxygenase reactions mol<sup>-1</sup>  
150 <sup>1</sup>],  $p$  is the number of electron equivalents (eeq) per mole substrate [eeq mol<sup>-1</sup>].

$\Delta G^{0'}_s$  is given by the Gibbs free energy associated with the transformation of the carbon source (which is often also the electron donor) to an intermediate and the subsequent synthesis of cell material from this intermediate

$$\Delta G^{0'}_s = \frac{\Delta G^{0'}_{fa} - \Delta G^{0'}_d}{\epsilon^m} + \frac{\Delta G^{0'}_{in} - \Delta G^{0'}_{fa}}{\epsilon^n} + \frac{\Delta G^{0'}_{pc}}{\epsilon} \quad (S14)$$

where  $\Delta G^{0'}_{fa}$  is the Gibbs free energy of the half-reaction of formaldehyde ( $\Delta G^{0'}_{fa} = 46.53 \text{ kJ eeq}^{-1}$ ),  $\Delta G^{0'}_{in}$  is the intermediate (assumed to be acetyl-CoA, a main intermediate in cell synthesis ( $\Delta G^{0'}_{in} = 30.9 \text{ kJ eeq}^{-1}$ )).  $m$  is +1 when the electron donor is a C1 compound, else it is equal to  $n$ .  $n$  is equal to +1 if  $\Delta G^{0'}_{in} - \Delta G^{0'}_d > 0$  (energy is needed); the energy needed to drive the reaction from electron donor to intermediate has to be divided with  $\epsilon$  to take inefficiencies into account. If  $\Delta G^{0'}_{in} - \Delta G^{0'}_d < 0$   $n$  is -1 (energy is generated); the energy generated from the reaction is captured with efficiency  $\epsilon$ . The efficiency term in synthesis ( $\epsilon$ ) is normally taken to be similar to the energy capture efficiency despite not being proven to be identical [11, 12].

$\Delta G^{0'}_{pc}$  is the Gibbs free energy associated with the synthesis of cell material from the intermediate is given as

$$\Delta G^{0'}_{pc} = \frac{\Delta G^{0'}_{ATP}}{Y_{ATP} \times 0.9} \times \frac{M_C}{\gamma_C \sigma_C} \quad [\text{kJ eeq}^{-1}] \quad (S15)$$

where  $\gamma_C$  is the average degree of reductance of the carbon atoms in the cell [ $\text{eeq} (\text{mol carbon})^{-1}$ ],  $\sigma_C$  is the fraction of carbon in the cell [for a cell formula of  $\text{C}_5\text{H}_7\text{O}_2\text{N}$ :  $\gamma_C = 4 \text{ eeq mol}^{-1}$ ,  $\sigma_C = 0.531 \text{ g carbon (g cell)}^{-1}$ ],  $\Delta G^{0'}_{ATP}$  is as previously defined,  $Y_{ATP}$  is as previously defined and taken as a constant and equal to  $10.5 \text{ g cell dw (mol ATP)}^{-1}$ , and 0.9 is the percentage of organic material of a dry cell [ $\text{g organic material (g cell dw)}^{-1}$ ].



175 *S1.3 Expanded Thermodynamic True Yield Prediction Model (ETTYM)*

176 The Expanded Thermodynamic True Yield Model was presented in [4] and expanded in  
177 [11] to account for oxygenase reactions and pH.

178 To increase the accuracy for the yield prediction on C1 compounds and  
179 substrates with low degrees of reduction, the authors proposed to include a carbon and a  
180 nitrogen balance and as a result thereof reformulate the electron and energy balance  
181 originally proposed by McCarty [4, 11]. The nitrogen balance can be ignored if nitrogen  
182 is not limiting [11], hence, the yield can be calculated from an energy balance, carbon  
183 balance, and electron balance.

185 *Carbon balance*

186 The carbon balance describes that the carbon is either invested in cell synthesis or into  
187 other carbonaceous products. If the only other carbonaceous product is CO<sub>2</sub>, the  
188 equation states that either carbon is incorporated into cell mass or oxidised to CO<sub>2</sub>

189 
$$f_{cell} + \sum_i f_{CS}(i) = 1 \quad (S16)$$

191 where  $f_{cell}$  is the bacterial growth yield [g cell carbon (g substrate carbon)<sup>-1</sup>], and  $f_{CS}(i)$  is  
192 the yield of carbonaceous product  $i$ .

194 *Energy balance*

195 The energy balance describes the relationship between the Gibbs free energy of cell  
196 synthesis and the Gibbs free energy captured from oxidation and reduction of various  
197 electron donor-electron acceptor pairs. If there is only one electron donor-electron  
198 acceptor pair, the energy balance can be written as

$$f_{cell} \times \left( \frac{(\Delta G_{acetate} - \Delta G_{CS})}{K^m} + \frac{\frac{\Delta G_{ATP} \times M_{cell}}{Y_{ATP} \times 0.9}}{K} \right) = -K \times (g(1)\Delta G_a - f_{CO_2}\Delta G_d) \quad (S17)$$

where  $\Delta G_{acetate}$  is the Gibbs free energy of acetate reduction (= 106.3 kJ (mol C)<sup>-1</sup>) [11]  
 $\Delta G_{CS}$  is the Gibbs free energy of the carbon source,  $M_{cell}$  is the cell mass per mol carbon  
 (= 26.4 g (mol C)<sup>-1</sup>) with cell formula C<sub>5</sub>H<sub>10</sub>O<sub>3</sub>N,  $g(1)$  is the number of electrons gained  
 by the electron acceptor (electron equivalents (eeq) (mol carbon substrate)<sup>-1</sup>),  $f_{CO_2}$  is the  
 fraction of carbon oxidised to CO<sub>2</sub> (mol C (mol carbon substrate)<sup>-1</sup>),  $K$  is the energy  
 efficiency parameter associated with (i) energy capture and storage in ATP and (ii)  
 energy transfer from ATP to cell synthesis ( $K = 0.41$ ) [4]. While these two different  
 processes have different efficiencies, the values are close enough to lump them into one  
 parameter [4].  $m$  is +1 if  $(\Delta G_{acetate} - \Delta G_{CS}) > 0$  (energy is needed for transformation  
 of carbon source to acetate) else  $m = -1$  (energy is gained from transformation of carbon  
 source to acetate). The other parameters are as previously defined for TEEM2. All half-  
 reactions are written as reductions similarly to TEEM2 and MTB, although  $\Delta G_a$  is in kJ  
 eeq<sup>-1</sup>,  $\Delta G_d$  is in kJ (mol substrate carbon)<sup>-1</sup>.

#### Electron balance

The electron balance states that the electrons coming from the oxidation of the electron  
 donor(s) are equal to the electrons used for reducing electron acceptors: the terminal  
 electron acceptor, nitrogen source different from ammonium, electrons lost in  
 oxygenase reactions; electrons used to reduce or increase the degree of reduction of the  
 carbon source to the same level as the cell [4]

$$\sum_i f_{ED}(i) \times (\gamma_{ED} - \gamma(i)) = \sum_j g(j) \quad (S18)$$

223 where  $f_{ED}(i)$  is the fraction of electron donor going to oxidised product  $i$ ,  $\gamma_{ED}$  is the  
224 degree of reduction of the electron donor,  $\gamma(i)$  is the degree of reduction of oxidised  
225 product  $i$ ,  $g(j)$  is the number of electrons sent to electron acceptor  $j$ .

226 If the electron donor is also the carbon source and the electrons are used only for  
227 reduction of the terminal electron acceptor and synthesis of new cell material, the  
228 equation reduces to (remember,  $\gamma_{CO_2} = 0$ )

229 
$$f_{CO_2} \times (\gamma_s - \gamma_{CO_2}) + f_{cell}(\gamma_s - \gamma_x) = f_{CO_2} \times \gamma_s + f_{cell}(\gamma_s - \gamma_x) = g(1) \text{ (S19)}$$

230 where  $\gamma_s$  is the degree of reduction of the electron donor (carbon source), and  $\gamma_x$  is the  
231 degree of reduction of the cell.

Table S1. Balanced half-reactions as reductions of the simple carbon substrates and their associated Gibbs free energy of the half-reaction ( $\Delta G_r^\circ$ ) in  $\text{kJ mol}^{-1}$  and  $\text{kJ (electron equivalent (eeq))}^{-1}$  at standard state conditions, except for  $\text{H}^+$  ( $=10^{-7} \text{ M}$ ).

Electron donor	Half-reaction	$\Delta G_r^\circ$	
		in $\text{kJ mol}^{-1}$	$\text{kJ eeq}^{-1}$
Acetate	$2 \text{ CO}_2 + 7 \text{ H}^+ + 8 \text{ e}^- \rightleftharpoons \text{C}_2\text{H}_3\text{O}_2^- + 2 \text{ H}_2\text{O}$	223.4	27.9
Citrate	$6 \text{ CO}_2 + 15 \text{ H}^+ + 18 \text{ e}^- \rightleftharpoons \text{C}_6\text{H}_5\text{O}_7^{3-} + 5 \text{ H}_2\text{O}$	608.1	33.8
Formaldehyde	$\text{CO}_2 + 4 \text{ H}^+ + 4 \text{ e}^- \rightleftharpoons \text{CH}_2\text{O} + \text{H}_2\text{O}$	185.8	46.5
Formate	$\text{CO}_2 + \text{H}^+ + 2 \text{ e}^- \rightleftharpoons \text{CHO}_2^-$	82.8	41.4
a-D-Fructose	$6 \text{ CO}_2 + 24 \text{ H}^+ + 24 \text{ e}^- \rightleftharpoons \text{C}_6\text{H}_{12}\text{O}_6 + 6 \text{ H}_2\text{O}$	982.7	40.9
Fumaric acid	$4 \text{ CO}_2 + 12 \text{ H}^+ + 12 \text{ e}^- \rightleftharpoons \text{C}_4\text{H}_4\text{O}_4 + 4 \text{ H}_2\text{O}$	458.8	38.2
a-D-Galactose	$6 \text{ CO}_2 + 24 \text{ H}^+ + 24 \text{ e}^- \rightleftharpoons \text{C}_6\text{H}_{12}\text{O}_6 + 6 \text{ H}_2\text{O}$	974.6	40.6
Gluconate	$12 \text{ CO}_2 + 42 \text{ H}^+ + 44 \text{ e}^- \rightleftharpoons \text{C}_{12}\text{H}_{22}\text{O}_{14}^{2-} + 10 \text{ H}_2\text{O}$	3151	71.6
a-D-Glucose	$6 \text{ CO}_2 + 24 \text{ H}^+ + 24 \text{ e}^- \rightleftharpoons \text{C}_6\text{H}_{12}\text{O}_6 + 6 \text{ H}_2\text{O}$	980.9	40.9
Glycerol	$3 \text{ CO}_2 + 14 \text{ H}^+ + 14 \text{ e}^- \rightleftharpoons \text{C}_3\text{H}_8\text{O}_3 + 3 \text{ H}_2\text{O}$	540.3	38.6
Glycine	$2 \text{ CO}_2 + \text{NH}_3 + 6 \text{ H}^+ + 6 \text{ e}^- \rightleftharpoons \text{C}_2\text{H}_5\text{NO}_2 + 2 \text{ H}_2\text{O}$	208.7	34.8
Glyoxylate	$2 \text{ CO}_2 + 3 \text{ H}^+ + 4 \text{ e}^- \rightleftharpoons \text{C}_2\text{HO}_3^- + \text{H}_2\text{O}$	210.9	52.7
Lactate	$3 \text{ CO}_2 + 11 \text{ H}^+ + 12 \text{ e}^- \rightleftharpoons \text{C}_3\text{H}_5\text{O}_3^- + 3 \text{ H}_2\text{O}$	391.4	32.6
a-Lactose	$12 \text{ CO}_2 + 48 \text{ H}^+ + 48 \text{ e}^- \rightleftharpoons \text{C}_{12}\text{H}_{22}\text{O}_{11} + 13 \text{ H}_2\text{O}$	2044	42.6
Malate	$4 \text{ CO}_2 + 10 \text{ H}^+ + 12 \text{ e}^- \rightleftharpoons \text{C}_4\text{H}_4\text{O}_5^{2-} + 3 \text{ H}_2\text{O}$	418.2	34.9
Malonate	$3 \text{ CO}_2 + 6 \text{ H}^+ + 8 \text{ e}^- \rightleftharpoons \text{C}_3\text{H}_2\text{O}_4^{2-} + 2 \text{ H}_2\text{O}$	269.4	33.7
Mannitol	$6 \text{ CO}_2 + 26 \text{ H}^+ + 26 \text{ e}^- \rightleftharpoons \text{C}_6\text{H}_{14}\text{O}_6 + 6 \text{ H}_2\text{O}$	1035	39.8
Oxalate	$2 \text{ CO}_2 + 2 \text{ e}^- \rightleftharpoons \text{C}_2\text{O}_4^{2-}$	114.0	57.0
Phenol	$6 \text{ CO}_2 + 28 \text{ H}^+ + 28 \text{ e}^- \rightleftharpoons \text{C}_6\text{H}_6\text{O} + 11 \text{ H}_2\text{O}$	799.1	28.5
Phenylacetic acid	$8 \text{ CO}_2 + 36 \text{ H}^+ + 36 \text{ e}^- \rightleftharpoons \text{C}_8\text{H}_8\text{O}_2 + 14 \text{ H}_2\text{O}$	1131	31.4
Propionate	$3 \text{ CO}_2 + 13 \text{ H}^+ + 14 \text{ e}^- \rightleftharpoons \text{C}_3\text{H}_5\text{O}_2^- + 4 \text{ H}_2\text{O}$	390.7	27.9
Pyruvate	$3 \text{ CO}_2 + 9 \text{ H}^+ + 10 \text{ e}^- \rightleftharpoons \text{C}_3\text{H}_3\text{O}_3^- + 3 \text{ H}_2\text{O}$	354.8	35.5
Sorbitol	$6 \text{ CO}_2 + 26 \text{ H}^+ + 26 \text{ e}^- \rightleftharpoons \text{C}_6\text{H}_{14}\text{O}_6 + 6 \text{ H}_2\text{O}$	1035	39.8

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Succinate	$4\text{ CO}_2 + 12\text{ H}^+ + 14\text{ e}^- \rightleftharpoons \text{C}_4\text{H}_4\text{O}_4^- + 4\text{ H}_2\text{O}$	415.7	29.7
Tartrate	$4\text{ CO}_2 + 8\text{ H}^+ + 10\text{ e}^- \rightleftharpoons \text{C}_4\text{H}_4\text{O}_6^{2-} + 2\text{ H}_2\text{O}$	583.9	58.4
Xylose	$5\text{ CO}_2 + 20\text{ H}^+ + 20\text{ e}^- \rightleftharpoons \text{C}_5\text{H}_{10}\text{O}_5 + 5\text{ H}_2\text{O}$	1231	61.5

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Table S2. Gibbs free energy of formation in  $\text{kJ mol}^{-1}$  ( $\Delta G_f^\circ$ ), number of carbon-hydrogen bonds,  $Y_{ATP}$  in g dry weight  $(\text{mol ATP})^{-1}$ , number of (putative) oxygenase reactions ( $t^{oxy}$ ), chemical structure, and degree of reductance ( $\gamma_s$ ) of the simple carbon substrates used in the comparison.

Name	Structure	$\Delta G_f^\circ$	C-H bonds	$Y_{ATP}$	$t_{oxy}$	$\gamma_s$	Reference
Acetate	$\text{C}_2\text{H}_3\text{O}_2^-$	-369.4	3	5	0	4	[13]
Citrate	$\text{C}_6\text{H}_5\text{O}_7^{3-}$	-1168.3	4	5	0	3	[13]
Formaldehyde	$\text{CH}_2\text{O}$	-130.5	2	5	0	4	[13]
Formate	$\text{CHO}_2^-$	-351.0	1	5	0	2	[13]
a-D-Fructose	$\text{C}_6\text{H}_{12}\text{O}_6$	-915.4	7	10	0	4	[13]
Fumaric acid	$\text{C}_4\text{H}_4\text{O}_4$	-647.1	2	5	0	3	[13]
a-D-Galactose	$\text{C}_6\text{H}_{12}\text{O}_6$	-923.5	7	10	0	4	[13]
Gluconate	$\text{C}_{12}\text{H}_{22}\text{O}_{14}^{2-}$	-880.2	12	10	0	3.7	[14]
a-D-Glucose	$\text{C}_6\text{H}_{12}\text{O}_6$	-917.2	7	10	0	4	[13]
Glycerol	$\text{C}_3\text{H}_8\text{O}_3$	-488.5	5	7.5	0	4.7	[13]
Glycine	$\text{C}_2\text{H}_5\text{NO}_2$	-370.8	2	5	0	3	[13]
Glyoxylate	$\text{C}_2\text{HO}_3^-$	-459.6	1	5	0	2	[13]
Lactate	$\text{C}_3\text{H}_5\text{O}_3^-$	-517.8	4	5	0	4	[13]
a-Lactose	$\text{C}_{12}\text{H}_{22}\text{O}_{11}$	-1515.2	14	5	0	4	[13]
Malate	$\text{C}_4\text{H}_4\text{O}_5^{2-}$	-845.1	3	5	0	3	[13]
Malonate	$\text{C}_3\text{H}_2\text{O}_4^{2-}$	-677.6	2	5	0	2.7	[14]
Mannitol	$\text{C}_6\text{H}_{14}\text{O}_6$	-942.6	8	5	0	4.3	[13]
Oxalate	$\text{C}_2\text{O}_4^{2-}$	-674.0	0	5	0	1	[13]
Phenol	$\text{C}_6\text{H}_6\text{O}$	-72.8	5	5	2	4.7	[14]
Phenylacetic acid	$\text{C}_8\text{H}_8\text{O}_2$	-136.9	7	5	2	4.5	[14]
Propionate	$\text{C}_3\text{H}_5\text{O}_2^-$	-361.1	5	5	0	4.7	[13]
Pyruvate	$\text{C}_3\text{H}_3\text{O}_3^-$	-474.6	3	5	0	3.3	[13]
Sorbitol	$\text{C}_6\text{H}_{14}\text{O}_6$	-942.7	8	5	0	4.3	[13]
Succinate	$\text{C}_4\text{H}_4\text{O}_4^{2-}$	-690.2	4	5	0	3.5	[13]
Tartrate	$\text{C}_4\text{H}_4\text{O}_6^{2-}$	-747.5	2	5	0	2.5	[14]

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Xylose	C <sub>5</sub> H <sub>10</sub> O <sub>5</sub>	-350.9	6	5	0	4	[14] 239
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Table S3. The relative deviation from the estimated bacterial growth yield estimated under aerobic conditions when assuming a Gibbs free energy of formation of  $0 \text{ kJ mol}^{-1}$ . The deviation is computed relative to the yield estimated with the Gibbs free energy of formation reported in Table 3. The mean absolute deviation is computed. MTB has the lowest mean average deviation.

	MTB	TEEM2	ETTYM
2,4-D	5%	7%	7%
2,4-DB	1%	1%	1%
Acetamiprid	-10%	-27%	-21%
Acetochlor	1%	2%	2%
Alachlor	-4%	-8%	-8%
Anthracene	-7%	-25%	-9%
Atrazine	-11%	-27%	-22%
Azoxystrobin	-3%	-5%	-5%
Benalaxyl	-4%	-8%	-7%
Benzene	-3%	-15%	-4%
Benzoate	2%	4%	3%
Bifenazate	-6%	-10%	-10%
Carbofuran	3%	5%	4%
Chlorothalonil		-6%	-8%
Chlorpropham	-7%	-11%	-11%
Cypermethrin	-4%	-6%	-6%
Daminozide	0%	0%	0%
DDT	-6%	-8%	-8%
Desmedipham	-8%	-12%	-12%
Dicamba	5%	7%	7%
Ethylenediaminetetraacetate (EDTA)	18%	39%	30%
Famoxadone	-6%	-9%	-8%
Glyphosate	20%	52%	52%
Iprodione	-5%	-7%	-7%

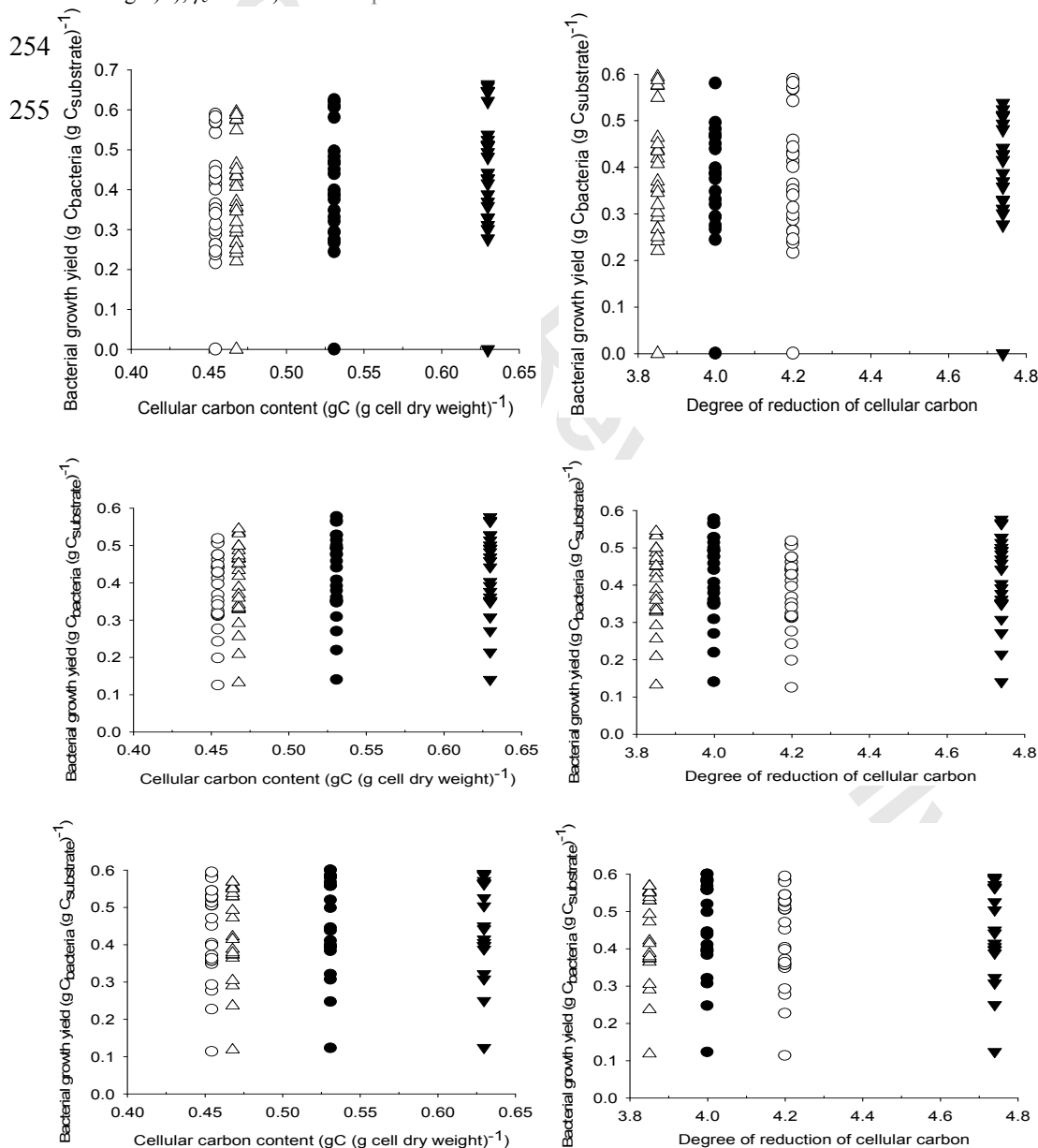


MCPA	2%	2%	2%
MCPB	0%	0%	0%
Mecoprop (MCP)	0%	0%	0%
Metalaxyl-M	-3%	-5%	-5%
Metamitron	-7%	-18%	-14%
Milbemectin	-3%	-6%	-6%
Naphthalene	-6%	-25%	-10%
Nitrilotriacetate (NTA)	23%	49%	44%
Paraquat	-8%	-27%	-14%
Pendimethalin	-7%	-13%	-13%
Phenanthrene	-7%	-25%	-9%
Phenmedipham	-8%	-11%	-12%
Propyzamide	-4%	-6%	-7%
Pymetrozine	-12%	-27%	-24%
Pyrene	-3%	-12%	-5%
Mean absolute deviation	6%±5%	14%±13	11%±11%

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Figure S1. The effect of cell formula on the yield estimate. **Top:** Microbial Turnover to Biomass (MTB) method; **middle:** Thermodynamic Electron Equivalent Method 2 (TEEM2); **bottom:** Expanded Thermodynamic True Yield Prediction Method (ETTYM). **Left:** The effect of the carbon content ( $\sigma_c$ ); and **right:** the effect of degree of reductance ( $\gamma_c$ ) are shown depending on the chosen cell formula. For ETTYM, the standard cell formula is  $C_5H_{10}O_3N$ ; for TEEM2 and MTB it is  $C_5H_7O_2N$ .  $\bullet$ :  $C_5H_7O_2N$  ( $\sigma_c = 0.53$  gC (g cell dry weight) $^{-1}$ ,  $\gamma_c = 4$ );  $\circ$ :  $C_5H_{10}O_3N$  ( $\sigma_c = 0.45$  gC (g cell dry weight) $^{-1}$ ,  $\gamma_c = 4.2$ );  $\blacktriangledown$ :  $C_5H_{8.33}O_{0.8}N$  ( $\sigma_c = 0.63$  gC (g cell dry weight) $^{-1}$ ,  $\gamma_c = 4.74$ );  $\Delta$ :  $C_{4.1}H_{6.8}O_{2.2}N$  ( $\sigma_c = 0.47$  gC (g cell dry weight) $^{-1}$ ,  $\gamma_c = 3.85$ ). The compounds used can be found in Table S2.



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257 **References**

- 258 [1] S. Trapp, A. Libonati Brock, K. Nowak and M. Kästner, *Prediction of the formation*  
259 *of biogenic non-extractable residues during degradation of environmental*  
260 *chemicals from biomass yields*, Environ. Sci. Technol. (2017), *Submitted for*  
261 *publication*
- 262 [2] G. Diekert, Grundmechanismen des Stoffwechsels und der Energiegewinnung, in  
263 *Umweltbiotechnologie*, J.C.G. Ottow and W. Bidlingmaier, eds., Fischer Verlag,  
264 Stuttgart, Germany, 1997; pp. 1-38.
- 265 [3] K. Burton, *Energy of adenosine triphosphate*. Nature 181 (1958), pp. 1594-1595.
- 266 [4] J. Xiao, and J.M. VanBriesen, *Expanded thermodynamic model for microbial true*  
267 *yield prediction*, Biotechnol. Bioeng 93 (2006), pp. 110–121.
- 268 [5] J.J. Heijnen, and J.P. Dijken, *In search of a thermodynamic description of biomass*  
269 *yields for the chemotrophic growth of microorgansims*, Biotechnol Bioeng 39  
270 (1992), pp. 833-852
- 271 [6] D. W. Tempest and O.M. Neijssel, *The Status of  $Y_{atp}$  and Maintenance Energy as*  
272 *Biologically Interpretable Phenomena*. Ann. Rev. Microbiol., 38 (1984), pp. 459–  
273 486.
- 274 [7] P.L. McCarty, *Thermodynamic Electron Equivalents Model for bacterial yield*  
275 *prediction: Modifications and comparative evaluations*, Biotechnol. Bioeng 97  
276 (2007), pp. 377–388.
- 277 [8] P.L. McCarty, *Thermodynamics of biological synthesis and growth*, Intl. J. Air  
278 Water Poll. 9 (1965), pp. 621-639.
- 279 [9] B.E. Rittmann, P.L. McCarty, *Environmental biotechnology: principles and*  
280 *applications*. McGraw-Hill: New York, NY, 2001.
- 281 [10] J.M. VanBriesen, *Thermodynamic yield predictions for biodegradation through*  
282 *oxygenase activation reactions*, Biodegradation 12 (2001), pp. 265-281.
- 283 [11] J. Xiao, and J.M. VanBriesen, *Expanded thermodynamic true yield prediction*  
284 *model: adjustments and limitations*, Biodegradation 19 (2008), pp. 99–127.
- 285 [12] J.M. VanBriesen, *Evaluation of methods to predict bacterial yield using*  
286 *thermodynamics*, Biodegradation 13 (2002), pp. 171–190.
- 287 [13] R.K. Thauer, K. Jungermann, and K. Decker, *Energy conservation in chemotrophic*  
288 *anaerobic bacteria*, Bacteriol. Rev. 41 (1977), pp. 100–180.

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3 [14] A. Flamholz, E. Noor, A. Bar-Even, and R. Milo, *eQuilibrator - the biochemical*  
4 *thermodynamics calculator*, Nucleic Acids Res. 40 (2012), pp. 770-775.  
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